 Aquatic Effects Monitoring Plan

## Juvenile Lake Sturgeon Population Monitoring Report

AEMP-2019-06


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# KEEYASK GENERATION PROJECT 

## AQUATIC EFFECTS MONITORING PLAN

REPORT \#AEMP-2019-06

# JUVENILE LAKE STURGEON POPULATION MONITORING, FALL 2018: YEAR 5 CONSTRUCTION 

Prepared for

Manitoba Hydro

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## SUMMARY

## Background

The Keeyask Hydropower Limited Partnership (KHLP) was required to prepare a plan to monitor the effects of construction and operation of the Keeyask Generating Station (GS) on the environment. Besides measuring the accuracy of the predictions made and actual effects of the GS on the environment, monitoring results will provide information on how construction and operation of the GS will affect the environment and if more needs to be done to reduce harmful effects.

Construction of the Keeyask GS began in mid-July 2014 with the construction of cofferdams that blocked flow in the north and central channels of Gull Rapids (see instream structures map below). During the winter of 2015/2016, the Spillway Cofferdam, which partially blocks the south channel, was constructed. Beginning late in 2016 and continuing in 2017, the Tailrace Cofferdam was constructed. Work was completed in fall 2017 with the exception of an opening that was left to allow fish movement into and out of the cofferdam over the 2017/18 winter. This opening was closed in spring 2018, and the area was dewatered. The spillway was commissioned in August 2018. The South Dam Cofferdam was completed in fall 2018, blocking the channel and forcing the entire flow of the river through the spillway.

Lake Sturgeon were identified as one of the key species for monitoring. They were chosen because they are culturally important to local people, the local populations have been previously impacted, and construction and operation of the GS will change or negatively impact important habitat. The plan to monitor the impacts of GS construction and operation on sturgeon includes several types of studies:

- Estimating the number of adults;
- Estimating the number and growth of juveniles (less than 800 millimetres [mm] in length);
- Identifying spawning locations and numbers of spawning fish; and
- Movement studies to record seasonal habitat use, long distance movements, and movements past barriers (i.e., over GSs or Gull Rapids).

The mitigation and offsetting plan for Lake Sturgeon included a commitment to a long-term stocking program. This plan addressed the loss of spawning habitat at Gull Rapids during the construction and initial years of operation (i.e., before the constructed spawning habitat is fully effective) by releasing young sturgeon into Stephens Lake. Stocking will also support the recovery of the sturgeon populations in Gull Lake, Stephens Lake, and the Upper Split Lake Area. Stocking began in 2014, with locations alternated between years (future Keeyask reservoir and Stephens Lake are stocked with fish born in even years, Burntwood River is
stocked with fish born in odd years) and its effectiveness is assessed through juvenile population monitoring.

This report presents results of juvenile Lake Sturgeon population monitoring conducted during fall 2018. Data from juvenile populations in the study area have been collected intermittently since 2007 and the juvenile population monitoring study was conducted for the first time in 2014. The plan is to conduct juvenile population monitoring annually until 2044. Each year, sampling will be conducted using the same capture methods, so that results can be compared between different years and trends can be seen.


Satellite Imagery - October 12th, 2018

Map illustrating instream structures at the Keeyask Generating Station site, October 2018.

## Why is the study being done?

Juvenile Lake Sturgeon population monitoring is being done to answer several questions:
Does recruitment of wild sturgeon occur upstream and/or downstream of the GS during construction?

This question is important because if no young sturgeon are born during the seven year construction period, then in the future fewer adult sturgeon will be reproducing.

Is there a change in condition factor and growth of juvenile sturgeon during construction?
This question is important because if sturgeon become fatter or skinnier than they used to be, then something is changing in their environment. If the condition of juveniles decreases, it can also mean that stocking is adding too many fish to the environment and they cannot find enough food. In that case, the stocking plan will be adjusted.

What is the survival rate of stocked sturgeon?
This question is important because if the survival rate is high then the number of fish stocked may be reduced. If the survival rate is low, then the stocking plan would be adjusted (e.g., may change time or location of release).

What is the proportion of hatchery-reared to wild recruits within a birth year (i.e., how successful is the stocking program)?

The answer to this question will also tell us about the effectiveness of the stocking program.


Juvenile (left) and young-of-the-year (right) Lake Sturgeon.

## What was done?

Sampling was conducted in the Upper Split Lake Area (including the Burntwood River and Split Lake), the future Keeyask reservoir (the Nelson River between Clark Lake and Gull Rapids), and Stephens Lake in the fall of 2018. Gill nets were used to catch juvenile sturgeon, defined as those that are less than 800 mm in length. The gill nets were set in deep water habitats preferred by juveniles. When a fish was caught, it was measured and weighed. If the fish was

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not already tagged, then two different tags were applied: an external (Floy ${ }^{\circledR}$ ) tag and a small PIT tag to make sure the fish is identifiable if one tag is lost. If the captured fish had already been tagged, then the tag numbers were recorded before the fish was released. Tagging and recapturing fish makes it possible to determine how much a fish grew or the distance they moved. It also makes it possible to estimate how many sturgeon are in a population. An ageing structure (a small piece of fin) was also collected to determine the year that the fish was born.


Captured juvenile Lake Sturgeon in a fish tub (left); measuring (middle); and weighing (right) a Lake Sturgeon after capture.

## What was found?

A total of 68 Lake Sturgeon were captured in the Upper Split Lake Area: 11 in the Burntwood River (all juveniles) and 57 in Split Lake ( 52 juveniles and 5 adults). In the Burntwood River, the presence of sturgeon born every year between 1999 and 2017 since sampling began shows that recruitment is occurring relatively consistently but at low levels. Sturgeon born in 2013 accounted for a large number of the fish caught in Split Lake. No Lake Sturgeon born in 2018 (called young-of-the-year [YOY]) were captured in either area. Of the 68 sturgeon caught, four were wild fish tagged in a previous year and recaptured in 2018: one in the Burntwood River and three in Split Lake. All four fish were recaptured in the waterbody where they were originally tagged. One fish captured in Split Lake was raised at the Grand Rapids hatchery and released in the Burntwood River as a one-year-old in 2014.

A total of 150 Lake Sturgeon (143 juveniles and 7 adults) were captured in the future Keeyask reservoir. Fish born in 2008 continued to be prominent in the catch, as well as fish from the 2016 cohort. One YOY sturgeon was captured, showing that sturgeon successfully reproduced in 2018 (no stocking of YOY took place in this area in 2018 prior to sampling). Eighteen of the 150 sturgeon had been tagged in a previous year (between 2006 and 2017), and an additional 17 were tagged hatchery-reared sturgeon released as one-year-olds in 2015, 2017, and 2018. These fish made up a large proportion of the 2014 (62\%) and 2016 (21\%) year classes. One of the captured hatchery fish was released 128 km upstream in the Burntwood River in 2018. This is the fourth fish captured in the future Keeyask reservoir that was originally stocked into the Burntwood River since 2014. Analysis of growth between hatchery and wild caught fish showed that young hatchery fish are longer and heavier than wild fish of the same cohort. However, growth of hatchery fish appears to slow around age-4.

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In Stephens Lake, 74 Lake Sturgeon (71 juvenile and 3 adult) were captured. The majority of the captures were sturgeon born in 2015 and represent wild fish spawned at Gull Rapids (as no 2015 cohort fish were stocked into Stephens Lake). No YOY (fish born in 2018) were captured in Stephens Lake. Eleven sturgeon tagged in a previous year were recaptured, as well as 17 hatchery-reared sturgeon (released as one-year olds in 2015 and 2017). Three of the hatcheryreared sturgeon were stocked in the future Keeyask reservoir while the remaining 14 were stocked in Stephens Lake. Hatchery fish made up a large proportion of the 2014 (77\%) and 2016 (64\%) year-classes. As in the future Keeyask reservoir, stocked hatchery fish were longer and heavier than wild fish of the same cohort, but showed slowed growth around age-4.

A computer model was used to generate estimates of population size and survival for wild juvenile Lake Sturgeon in the future Keeyask reservoir and Stephens Lake. Previously, not enough fish had been captured for the model to work, so 2018 was the first year estimates were calculated. In 2018, the future Keeyask reservoir population was estimated at 4,133 fish. Survival in this area was $77 \%$. The Stephens Lake population in 2018 was estimated at 1,101 individuals, and survival was estimated at $88 \%$.

A different model was also used to generate survival estimates for hatchery-reared fish stocked in the future Keeyask reservoir and Stephens Lake. Only enough fish stocked in 2015 (born in 2014) were caught to generate an estimate. Survival of fish stocked in 2015 was estimated at $95 \%$ in the future Keeyask reservoir and $86 \%$ in Stephens Lake.

As more data are collected and added to the models, the population and survival estimates get more precise and accurate. This is especially true as 2018 was the first time estimates could be calculated. As more fish are recaptured, estimates will become more refined.

## What does it mean?

The capture of YOY sturgeon in 2018 shows that, like in 2015-2017, reproduction in the wild is occurring upstream of Gull Rapids during Keeyask construction. No YOY were captured in the Upper Split Lake Area or Stephens Lake but that does not mean reproduction was unsuccessful in 2018. Prior to 2018, sturgeon have been born in each year since construction started (20152017) and in each of the three areas (Upper Split Lake, the future Keeyask reservoir, and Stephens Lake).

There were some changes in growth rate and condition between fish caught before construction and fish caught during construction in both the future Keeyask reservoir and Stephens Lake. However, there was no clear pattern and growth curves were similar.

The capture of a large number of hatchery-reared sturgeon released as one-year-olds in the future Keeyask reservoir and Stephens Lake over the last two study years suggests the stocking program is having a positive effect on juvenile numbers in these areas. It demonstrates that stocked sturgeon are surviving in the wild and that they are growing after release. Very few hatchery-raised fish have been captured in the Burntwood River, with similar low numbers captured in Split Lake and the future Keeyask reservoir. This indicates that fish may not stay within the Burntwood River after stocking.

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## What will be done next?

Monitoring will continue each fall until 2044. Further monitoring will show whether construction is affecting the growth of juveniles in the future Keeyask reservoir and Stephens lakes and whether sturgeon continue to reproduce. Survival, growth, and population size of stocked and wild juveniles will continue to be analysed.

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### 1.0 INTRODUCTION

The Keeyask Generation Project (the Project) is a 695-megawatt (MW) hydroelectric generating station at Gull Rapids on the lower Nelson River in northern Manitoba. The Project is approximately 725 kilometres (km) northeast of Winnipeg, 35 km upstream of the existing Kettle Generating Station, where Gull Lake flows into Stephens Lake, 60 km east of the community of Split Lake, 180 km east-northeast of Thompson and 30 km west of Gillam (Map 1). Construction of the Project began in July 2014.

The Keeyask Generation Project: Response to EIS Guidelines, completed in June 2012, provides a summary of predicted effects and planned mitigation for the Project. Technical supporting information for the aquatic environment, including a description of the environmental setting, effects and mitigation, and a summary of proposed monitoring and follow-up programs, is provided in the Keeyask Generation Project Environmental Impact Statement: Aquatic Environment Supporting Volume (AE SV). As part of the licensing process for the Project, an Aquatic Effects Monitoring Plan (AEMP) was developed detailing the monitoring activities of various components of the aquatic environment, including the focus of this report, juvenile Lake Sturgeon populations, for the construction and operation phases of the Project.

For the purposes of this monitoring program, Lake Sturgeon that are 800 mm in fork length or longer are classified as adults and smaller sturgeon are considered juveniles. Although fish greater than 800 mm length may not yet be sexually mature and may not reach sexual maturity for some years, this length was used as the threshold to distinguish between juveniles and adults because the smallest mature fish captured to date has been 809 mm (captured in 2016 in Stephens Lake; Legge et al. 2017).

Juvenile population monitoring is a key component of the overall Lake Sturgeon monitoring program. The Project is predicted to affect sturgeon recruitment by altering spawning habitat at Gull and Birthday rapids. Stocking aims to assist the recovery of sturgeon populations in the Upper Split Lake Area (i.e., the Burntwood River and the Nelson River between the Kelsey GS and Split Lake) and in the future Keeyask reservoir and Stephens Lake and is a key component of the offsetting plan, with stocking locations alternating between years (future Keeyask reservoir and Stephens Lake are stocked with even-cohort years, Burntwood River is stocked with odd-cohort years). Results of juvenile population monitoring will determine the impact of the loss of spawning habitat earlier than would be possible using adult population monitoring data, allowing timely adaptive management and mitigation, if required. Results of juvenile population monitoring will also assist in assessing the effectiveness of stocking and identify whether changes to the stocking plan are required. Data collected during juvenile population monitoring will be used to measure population size and cohort strength, identify changes in condition factor, determine whether natural reproduction is occurring, assess the need for young-of-the-year (YOY) habitat creation, and determine whether stocked fish are surviving and growing.

Juvenile Lake Sturgeon studies have been conducted in Gull Lake (the future Keeyask reservoir) and Stephens Lake since 2007. Surveys were initiated in the Burntwood River in 2012
and in the Nelson River downstream of the Kelsey GS and in Split Lake in 2015. These studies have increased the understanding of YOY and juvenile abundance, distribution, habitat use, condition, size, and year-class strength (MacDonald 2009; Michaluk and MacDonald 2010; Henderson and Pisiak 2012; Henderson et al. 2011, 2013, 2015; Burnett et al. 2016, 2017, 2018). Results from the Burntwood River show that small numbers of juvenile Lake Sturgeon are spread amongst multiple cohorts, indicating that recruitment has occurred fairly consistently in the Burntwood River over the previous 10 years (Henderson and Pisiak 2012; Henderson et al. 2013, 2015; Burnett et al. 2017, 2018). Catches of juvenile Lake Sturgeon in the Nelson River downstream of the Kelsey GS and in Split Lake are low, but a range of ages is represented. In both Gull and Stephens lakes, recruitment has also occurred fairly consistently over the past ten years, but until recently the cohort-frequency distribution has been dominated by a single cohort produced in 2008 (Henderson et al. 2011, 2013, 2015; Henderson and Pisiak 2012; Burnett et al. 2017, 2018). It has been shown that the growth of Burntwood River Lake Sturgeon is slower than conspecifics captured in Gull and Stephens lakes (Henderson et al. 2013).

Lake Sturgeon stocking is being conducted using wild caught broodstock from the Burntwood River and from the future Keeyask reservoir. To maintain the genetic structure of each population, progeny from each broodstock location are released back into their respective rivers (i.e., Burntwood River progeny released back into the Burntwood River and future Keeyask reservoir progeny released back to the Nelson River in Gull and Stephens lakes). Stocking occurred for the first time in 2014 and has occurred annually since with a variety of life stages (larvae, fingerlings, yearlings) being released (Table 1; Klassen et al. 2017, 2018, 2019).

This report presents results from the 2018 juvenile population monitoring conducted in the Upper Split Lake Area, future Keeyask reservoir, and Stephens Lake. In 2018, the area downstream of the Kelsey GS was not sampled as part of the Upper Split Lake Area in order to focus on areas where hatchery fish stocked into the Burntwood River may be captured (i.e., the Burntwood River and Split Lake).

Juvenile monitoring is being conducted to address the following key questions relevant during the construction period, as described in the AEMP:

- Does recruitment of wild sturgeon occur upstream and/or downstream of the GS during construction?
- Is there a biologically meaningful (and statistically significant) change in condition factor and growth of juvenile sturgeon during construction?
- What is the survival rate of stocked sturgeon?
- What is the proportion of hatchery-reared to wild recruits within a cohort (i.e., how successful is the stocking program)?

Juvenile population monitoring data will be collected annually until 2044.

### 2.0 STUDY SETTING

Juvenile population monitoring in 2018 was conducted at three locations: 1) the Upper Split Lake Area (Burntwood River and Split Lake); 2) the future Keeyask reservoir (i.e., the reach of the Nelson River between the outlet of Clark Lake and Gull Rapids), and 3) Stephens Lake.

The Burntwood River flows in a north-easterly direction from First Rapids for approximately 35 km prior to emptying into the western arm of Split Lake (Maps 1 and 3). It is unknown if First Rapids represents a natural barrier to upstream fish passage; however, it is assumed to be under high flow conditions. Hard substrates predominate in the main channel, while loose, fine sediments and associated macrophyte growth occur in many off-current areas.

Split Lake is located at the confluence of the Burntwood and Nelson rivers (Map 1). Due to the large inflows from the Nelson and Burntwood rivers, the lake has detectable current in several locations. Split Lake has maximum and mean depths of 28.0 m and 3.9 m , respectively, at a water surface elevation of 167.0 m above sea level (ASL; Lawrence et al. 1999). The surface area of Split Lake was determined to be 26,100 ha (excluding islands), with a total shoreline length, including islands, of 940.0 km (Lawrence et al. 1999). The numerous islands in Split Lake represent 411.6 km of the total shoreline.

Birthday Rapids is located approximately 10 km downstream of Clark Lake and 30 km upstream of Gull Rapids (Map 1). The drop in elevation from the upstream to downstream side of Birthday Rapids is approximately 2 m . The 14 km reach of the Nelson River between Birthday Rapids and Gull Lake is characterized as a large, somewhat uniform channel with medium to high water velocities. There are a few large bays with reduced water velocity and a number of small tributaries that drain into the Nelson River.

Gull Lake is a section of the Nelson River where the river widens, with moderate to low water velocity. Gull Lake is herein defined as the reach of the Nelson River beginning approximately 17 km upstream of Gull Rapids and 14 km downstream of Birthday Rapids, where the river widens to the north into a bay around a large point of land (Maps 1 and 5), and extending to the downstream end of Caribou Island, approximately 3 km upstream of Gull Rapids. Gull Lake has three distinct basins, the first extending from the upstream end of the lake downstream approximately 6 km to a large island; the second extending from the large island to Morris Point (a constriction in the river immediately upstream of Caribou Island); and the third extending from Morris Point to the downstream end of Caribou Island.

Gull Rapids is located approximately 3 km downstream of Caribou Island on the Nelson River (Map 1). The rapids are approximately 2 km in length, and the river elevation drops approximately 11 m along its 2 km length. Two large islands and several small islands occur within the rapids, prior to the river narrowing; these features are within the Project footprint and have been substantially altered during construction (Map 2). A summary of construction activities is provided in Section 2.1.

Just below Gull Rapids, the Nelson River enters Stephens Lake (Maps 1 and 7). Stephens Lake was formed in 1971 by construction of the Kettle GS. Between Gull Rapids and Stephens Lake, there is an approximately 6 km long reach of the Nelson River that, although affected by water regulation at the Kettle GS, remains riverine habitat with moderate velocity. Construction has altered the flow distribution immediately downstream of Gull Rapids as all flow now passes via the south channel of Gull Rapids. In August 2018, flow was further constricted when the spillway was commissioned (see Section 2.1).

Construction of the Kettle GS flooded Moose Nose Lake (north arm) and several other small lakes that previously drained into the Nelson River, as well as the old channels of the Nelson River that now lie within the southern portion of the lake. Major tributaries of Stephens Lake include the North and South Moswakot rivers that enter the north arm of the lake. Looking Back Creek is a second order stream that drains into the north arm of Stephens Lake (Map 1). Kettle GS is located approximately 40 km downstream of Gull Rapids.

### 2.1 CONSTRUCTION Summary

Construction of the Keeyask GS began in mid-July 2014 with the construction of cofferdams in the north and central channels of Gull Rapids (Map 2). These cofferdams resulted in the dewatering of the north and central channels and the diversion of all flow to the south channel. Construction of the Spillway Cofferdam (SWCD), which extends into the south channel of Gull Rapids, was completed in 2015. The rock placement for the inner and outer groins of the Tailrace Cofferdam (TRCD) started in late 2016 and the impervious fill placement was completed in fall 2017. An opening was created to allow fish to move freely over the winter of 2017-2018. The opening was closed in spring 2018 and dewatering of the TRCD occurred in July, at which time a fish salvage was completed. In preparation for commissioning of the spillway, the SWCD was watered-up on both sides of the structure in June 2018. Removal of the SWCD started in early July and continued into August. The spillway was commissioned between August 3 and 7, 2018. Closing the south channel with the upstream South Dam Cofferdam (SDCD) commenced at the beginning of August and river closure was achieved on August 16. This closure and the work that continued to seal the cofferdam forced the entire river flow through the spillway. The downstream SDCD was completed in September and the area between the two cofferdams was dewatered, allowing for fish salvage to be completed by late September 2018. Work continued on the upstream SDCD until it was complete in late fall 2018.

### 2.2 Flows and Water Levels

From October 2017 to October 2018, Split Lake outflow ranged from about 2,800-4,000 m³/s. Flow typically fell in the range of about $3,000-3,500 \mathrm{~m}^{3} / \mathrm{s}$, which is near the historical annual median flow of approximately $3,300 \mathrm{~m}^{3} / \mathrm{s}$. Flow was generally higher during the 2017/2018 winter period, gradually declining from about $3,800 \mathrm{~m}^{3} / \mathrm{s}$ at the end of February 2018 to about
$2,800 \mathrm{~m}^{3} / \mathrm{s}$ by the beginning of May. From early May 2018 to the beginning of July, flow gradually increased to about $3,500 \mathrm{~m}^{3} / \mathrm{s}$ and remained at that level to the end of July. The flow subsequently declined to about $2,800 \mathrm{~m}^{3} / \mathrm{s}$ by the end of September. Water levels varied in conjunction with the flows, ranging from about 153.4-155.2 m ASL on Gull Lake.

### 3.0 METHODS

### 3.1 Gillnetting

A standardized sampling methodology has been developed for sampling juvenile sturgeon in Boreal Shield rivers using data sets collected from several populations in the Hudson Bay drainage basin (McDougall et al. 2014). This standardized methodology (described below) is being used to enable comparisons of cohort strength, abundance, growth, and condition among years. The gillnetting methods described below have been used to capture juvenile Lake Sturgeon during environmental studies related to the Keeyask Generation Project since 2008.

Gillnetting was conducted in the Upper Split Lake Area, the future Keeyask reservoir, and the upper 10 km of Stephens Lake. Two locations were sampled in the Upper Split Lake Area, the Burntwood River between First Rapids and Split Lake and Split Lake proper. In previous years, the Nelson River between the Kelsey GS and Split Lake was also sampled. This area was not sampled in 2018 in order to focus effort on the Burntwood River and Split Lake as a means to locate stocked fish. Sites in Split Lake were chosen based on depth (greater than 5 m ) to try and target juvenile Lake Sturgeon. Gill nets were composed of five panels of 1, 2, 3, 5, and 6" twisted nylon stretched mesh (25, 51, 76, 127, and 152 mm ). Each panel was 25 yards (yd) $(22.9 \mathrm{~m})$ long and $2.7 \mathrm{yd}(2.5 \mathrm{~m})$ deep. Mesh sizes were staggered in the order of $1,5,2$, 6 , and 3 " to capture small and large juveniles across the length of each gang.

Gill nets were set in deep-water habitats (average depth $=11.6 \mathrm{~m}$ ) since YOY and juvenile Lake Sturgeon have been found to prefer these areas in the Winnipeg, Burntwood, and Nelson rivers (Barth et al. 2009; Michaluk and MacDonald 2010; McDougall et al. 2013; Henderson et al. 2014). Each gillnet set was given a unique identification number, and net locations were recorded using a Garmin Etrex GPS receiver (Garmin International Inc., Olathe, KS). Water depth at each end of the net was measured using a PiranhaMax Series 150 Portable Sonar (Humminbird, Eufaula, AL). Water temperature was measured daily in each area using a handheld thermometer $\left( \pm 0.5^{\circ} \mathrm{C}\right)$. HOBO Water Temperature Pro data loggers $\left( \pm 0.2^{\circ} \mathrm{C}\right)$, set approximately 1 m off the substrate, were also used to log water temperature at 6-hour intervals in Gull and Stephens lakes. Gill nets were checked approximately every 24 hours, weather permitting. For comparability among years, similar gillnetting locations were used during juvenile monitoring programs conducted from 2014 to 2018. However, some sites have changed between years depending on water levels and flows. Locations and site-specific physical measurements collected at gillnetting sites in 2018 are found in Appendix 1.

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### 3.2 Biological Sampling

All fish captured were counted by species and location. Lake Sturgeon were measured for fork length (FL; $\pm 1 \mathrm{~mm}$ ), total length ( $\pm 1 \mathrm{~mm}$ ), and weight ( $\pm 5 \mathrm{~g}$ using a digital scale, or nearest 25 g for fish greater than $4,000 \mathrm{~g}$ ).

For age analysis, the first fin ray of the left pectoral fin was removed immediately adjacent its articulation from each juvenile Lake Sturgeon captured for the first time. In cases where Lake Sturgeon had been previously aged, the first fin ray of the right pectoral fin was collected. If fish appeared to have been aged twice before or had deformed pectoral fins, ageing structures were not collected. All collected fin rays were placed in individually numbered envelopes, air dried, and brought back to the North/South Consultants Inc. laboratory for ageing (Section 3.4).

Small samples ( $1-2 \mathrm{~cm}^{2}$ ) were removed from the left pelvic fin of each Lake Sturgeon and preserved in $95 \%$ Biological Grade Ethanol for potential future genetic analysis.

Ageing structures and genetics samples were not taken from YOY fish due to concerns of harming the small fish. Ages were inferred based on size (i.e., fish smaller than 150 mm FL ).

### 3.3 TAGGING

Lake Sturgeon greater than 250 mm FL were marked with individually numbered external Floy-GD-94 T-bar (FT) anchor tags (Floy-tag Inc., Seattle, WA). Floy-tags were inserted into the base of the dorsal fin using a Dennison Mark II tagging gun (Avery Dennison Corporation, Pasadena, CA).

Uniquely numbered Passive Integrated Transponder (PIT) tags from Oregon RFID (Oregon RFID Ltd., Portland, OR) were also used to mark Lake Sturgeon. Those measuring greater than 250 mm FL received 12 mm HDX tags ( $12.0 \mathrm{~mm} \times 2.12 \mathrm{~mm} ; 0.1 \mathrm{~g}$ ) and those measuring less than 250 mm FL (smallest fish tagged was 99 mm ) received 8 mm FDX-B tags ( $8.0 \mathrm{~mm} x$ $1.4 \mathrm{~mm} ; 0.027 \mathrm{~g}$ ). Each Lake Sturgeon was scanned for an existing PIT tag using an Agrident APR 350 Reader (Agrident Ltd. Steinkippenstrasse, Germany). For each untagged fish, a PIT tag was injected under the third dorsal scute using an Oregon RFID tag injector needle, dipped in Polysporin® to minimize the risk of infection. Tags were injected parallel to the horizontal axis of the fish, into muscle tissue (not the body cavity). Following implantation or upon recapture, the tags were logged, and the last six digits manually recorded. Injector needles were sterilized in boiling water prior to the start of sampling and again upon sampling completion.

### 3.4 AgEING ANALYSIS

Lake Sturgeon fin rays were hardened in an epoxy resin (Cold Cure) and two 0.7 mm fin sections were cut distally within 5 mm of the articulation using a Struers Minitom (Struers Inc.

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Cleveland, OH ) low-speed sectioning saw. Fin sections were mounted on glass slides using Cytoseal-60 (Thermo Scientific, Waltham, MA) and viewed at five times magnification under a compound microscope. Annuli (growth rings) were counted by three experienced readers (independently), without prior knowledge of fish length or weight, or ages assigned by other readers. If readers assigned different ages to a fish, either the modal age or the median age was chosen. The rate of three-reader agreement was calculated in percent (percentage). Examples of Lake Sturgeon ageing structures are provided in Appendix 3.

Lake Sturgeon ageing structures exhibit well-defined banding patterns characteristic of repeated summer (fast-growth) and winter (slow/non-growth) periods (McDougall and Pisiak 2014; Appendix A3-1). Ageing structures from hatchery-reared Lake Sturgeon have different banding patterns that complicate the ageing process. During the winter period, when the water temperature remains close to $0^{\circ} \mathrm{C}$ and growth slows, wild Lake Sturgeon develop a well-defined annulus that is easy to recognize. However, hatchery reared sturgeon are often held at considerably higher and more constant temperatures than those experienced by wild fish (McDougall and Pisiak 2014). The higher temperatures over the first winter of growth cause a weak or missing first annulus. To account for this, current ageing methods add one year to the ages of fish with a weak/absent first annuli (McDougall and Pisiak 2012, 2014; Burnett and McDougall 2015) (see photo Appendix A3-2).

In fish stocked at age-1, the weak annulus is often followed by the presence of a false annulus, not corresponding to slowed winter growth, but instead to stocking and the subsequent establishment period. The false annuli decrease ageing accuracy because they are difficult to distinguish from true annuli. As a result, hatchery fish younger than 3-years-old are often overaged by one year, as the false annulus is counted. Ageing accuracy has been found to increase with time spent in the wild (Burnett and McDougall 2015; McDougall and Nelson 2016). All ageone hatchery fish are tagged with PIT tags, so ages are known. Of the 35 known hatchery fish captured in 2018, all but one were over-aged by one year. These ages were decreased to their known age for analysis.

In cases where a fish is captured with a weak or missing first annulus but does not have a PIT tag, it cannot be determined whether the fish is hatchery-reared or wild. Weak annuli occasionally occur in wild fish, but more likely, these fish were stocked as yearlings and shed their PIT tags. Therefore, to account for this uncertainty, these fish were not included in analyses of hatchery-reared vs. wild fish but were treated as a separate group (i.e., suspected hatchery fish).

### 3.5 DATA ANALYSIS

As was done in previous years, data were analysed for all sizes of Lake Sturgeon captured (as opposed to only those measuring less than 800 mm FL). Mesh sizes used select for small Lake Sturgeon but larger fish are also captured; therefore, including all fish in the summary statistics ensures comparability among years.

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To better describe sampling locations, relative abundance (CPUE), and fish movements, each sampling area was divided into distinct geographical zones (Maps 3, 5, and 7).

Mean FL (mm), weight (g), and condition factor (K) were calculated for all Lake Sturgeon by location. In Stephens Lake and the future Keeyask reservoir, known hatchery and wild fish were presented separately. Lake Sturgeon not confirmed as being either hatchery or wild (based on ageing structure analysis; described in Section 3.4) were only included in totals. Condition factor was calculated based on the following equation (after Fulton 1911, in Ricker 1975):

$$
K=W /\left(L^{3} / 10^{5}\right)
$$

Where:

$$
\begin{aligned}
& \mathrm{W}=\text { round weight }(\mathrm{g}) ; \text { and } \\
& \mathrm{L}=\text { fork length }(\mathrm{mm}) .
\end{aligned}
$$

Ageing structures were only collected for fish measuring < 800 mm FL. Because fish approach this length by age nine, all age analyses were restricted to fish aged $0-9$ years as the full range of sizes for older fish would not be included in the sample.
To determine if growth and condition of juvenile sturgeon were affected by construction, mean length-at-age and mean condition factor-at-age were compared using t-tests and Mann-Whitney U-Tests among all fish captured pre-Project (i.e., 2008-2013) and those captured during construction (i.e., 2014-2018 ${ }^{1}$ ). Tagged hatchery fish and suspected hatchery fish were removed from these analyses. Mann-Whitney U-Tests and t-tests were used to compare mean fork length-at age, mean weight-at-age, and mean condition factor-at-age for known hatcheryreared and wild Lake Sturgeon. Statistical comparisons were only conducted where sample sizes were greater than eight individuals. Significance level was set at 0.05 ( $5 \%$ ).

A von Bertalanffy growth curve was generated from all age and length data collected during the study, to compare the growth of wild vs. hatchery-reared fish, as well as wild fish captured during baseline vs. construction for fish aged as nine years or less. The curve was calculated using the following equation:

$$
L=L_{\infty}\left(1-\mathrm{e}^{-k(t-t} 0\right)
$$

Where:
$t=$ age (years)
$\mathrm{t}_{0}=$ is the theoretical age at which FL is 0 ;
$L=$ is the fork length (mm) of the fish at age $t$;
$L_{\infty}=$ is the theoretical maximum TL that an individual in the population can attain; and
$\mathrm{k}=$ growth rate .

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Length-frequency distributions were plotted in 50 mm length class intervals (e.g., 300-349 mm) and length-weight regression equations were derived using least squares analysis on logarithmic transformations of fork lengths and weights according to the following relationship:

$$
\ln (W)=\ln (a)+\ln (L)^{*} b
$$

Where:

$$
\begin{aligned}
& \mathrm{W}=\text { weight }(\mathrm{g}) ; \\
& \mathrm{L}=\text { fork length }(\mathrm{mm}) ; \\
& \mathrm{a}=\mathrm{Y} \text {-intercept; and } \\
& \mathrm{b}=\text { slope of the regression line. }
\end{aligned}
$$

Cohort frequency distributions were plotted for each location.
Gillnetting hours (i.e., effort) was calculated as the number of sampling hours per 100 m of net set using the following equation:

Effort (hours) $=$ set duration $\times($ net length $/ 100 \mathrm{~m})$
Catch-per-unit-effort (CPUE) was calculated and expressed as the number of fish captured in 100 m of net per 24-h period using the following formula:

$$
\text { CPUE }=\sum \text { \# Lake Sturgeon } / \sum \text { Effort } \times 24 \text { h }
$$

Where: $\Sigma=$ sum of the number of fish or gillnetting hours at all sites.
CPUE was calculated by geographical zone for each study location and study year.
Hatchery-reared Lake Sturgeon are released as larvae, fingerlings, and yearlings. However, without additional analysis (genetics or isotopic signature in fin rays) fish can only be conclusively identified as hatchery-reared based on the presence of a PIT tag, which are exclusive to fish stocked at age-one. All fish not definitively identified as hatchery-reared (based on the presence of a PIT tag) were classified as "wild" in order to facilitate data analysis. The exceptions to this classification are fish that are suspected to be stocked yearlings that have lost their PIT tags based on size and ageing structure features (see Section 3.4), which were noted as "suspected hatchery fish". As the additional analysis (genetics or isotopic signature in fin rays) has not been undertaken, it cannot be determined if fish belonging to cohorts corresponding to stocking events of larvae or fingerlings (e.g., a YOY captured in the Burntwood River in 2017) originated from the hatchery or a natural spawning event.

### 3.6 Population Estimate

Mark-recapture population estimates have been calculated for the future Keeyask reservoir and Stephens Lake during the fall of eight different years (2010 and 2012-2018). Only wild Lake Sturgeon classified as juveniles (i.e., fork length less than 800 mm ) were included in the population estimate. Juvenile Lake Sturgeon in the future Keeyask reservoir and Stephens Lake

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have been tagged in studies since 2008. All data for the period 2008-2012 were collected annually as part of environmental studies related to the pre-Project environment, while data from 2014 until 2044 will be collected annually as part of monitoring studies related to the Keeyask GS Project.

The Jolly-Seber model (POPAN formulation; Arnason and Schwarz 2002), as implemented within MARK, was used to estimate the annual abundance of juvenile Lake Sturgeon in the future Keeyask reservoir and in Stephens Lake. Detailed methods can be found in Appendix 5. Estimates are reported as a mean with $95 \%$ confidence intervals (CI).

A Cormack-Jolly-Seber model was used to calculate a survival estimate for hatchery-reared juvenile Lake Sturgeon cohorts with a minimum number of recaptures $(\mathrm{n}=25)$ stocked in both the future Keeyask reservoir and Stephens Lake, using the probability of recapture in each year. To date, only yearlings stocked in 2015 (i.e., the 2014 cohort) have had sufficient recaptures to conduct this analysis. The survival estimate was used to calculate the number of individuals from the 2014 hatchery cohort in both the future Keeyask reservoir and Stephens Lake in each year.

### 4.0 RESULTS

Biological and tagging information for Lake Sturgeon captured in 2018 are provided in Appendix 2.

### 4.1 Upper Split Lake Area

Water temperature in the Upper Split Lake Area ranged from $16.0^{\circ} \mathrm{C}$ to $9.0^{\circ} \mathrm{C}$ over the course of the study (September 9 to 20, 2018; Appendix A1-1).

### 4.1.1 BURNTWOOD RIVER

Six fish species $(\mathrm{n}=49)$ were captured at 19 sites in the Burntwood River between September 12 and 17, 2018 (Tables 2 and 3; Map 3). Lake Sturgeon ( $n=11 ; 22.4 \%$ ) were the second most abundant species captured after Longnose Sucker ( $n=24 ; 49.0 \%$ ) (Table 3). Gillnet site data as well as biological and tagging information for all Lake Sturgeon captured in the Burntwood River are provided in Appendices A1-1 and A2-1. Eleven juvenile Lake Sturgeon were captured in 426.3 gillnet hours, producing an overall CPUE of 0.62 LKST/ 100 m net/24 h (Table 4). No Lake Sturgeon mortalities occurred during sampling. Gill nets were set in all three zones of the Burntwood River below First Rapids; six sites in BWR-A, five sites in BWR-B, and eight sites in BWR-C (Map 3). CPUE values by zone were as follows:

- $1.03 \mathrm{LKST} / 100 \mathrm{~m}$ net/24 h in Zone BWR-A;
- 0.44 LKST/100 m net/24 h in Zone BWR-B; and
- 0.40 LKST/100 m net/24 h in Zone BWR-C (Table 4).

Annual CPUE values for the Burntwood River catch since 2011 are presented in Table 5.

### 4.1.1.1 Year-Class Strength

Ageing structures were collected from ten of 11 juvenile Lake Sturgeon. Three-reader agreement for age assignment was $90 \%(\mathrm{n}=9)$. The modal age was used for one fish (age 8), aged one year lower or higher by a single reader. Aged Lake Sturgeon ranged from 1 to 19 years old, with the $2010(n=2$; age 8), $2015(n=2$; age 3), and $2016(n=2$; age 2) cohorts each accounting for $20 \%$ of the catch (Figure 1). No YOY (2018 cohort) Lake Sturgeon were captured. Several other cohorts (2000-2001, 2003-2009, 2011, and 2014) were not present in the catch. Cohort frequencies for all juvenile Lake Sturgeon captured in the Upper Split Lake Area from 2011 to 2018 are presented in Table 6. Every cohort between 1999 and 2017 has been present in the catch since studies began.

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### 4.1.1.2 Growth and Condition

Captured Lake Sturgeon had a:

- Mean FL of $455 \mathrm{~mm}(\mathrm{n}=11$; standard deviation [StDev] = 191 mm ; range 205-764 mm);
- Mean weight of $950 \mathrm{~g}(\mathrm{n}=11$; StDev = 1,043 g; range 25-3,000 g); and
- Mean condition factor of $0.62(\mathrm{n}=11$; StDev $=0.23$; range $0.20-0.87)$ (Table 7).

Lake Sturgeon in the 250-299 mm ( $\mathrm{n}=2$ ), $350-399 \mathrm{~mm}(\mathrm{n}=2)$, and $450-499 \mathrm{~mm}(\mathrm{n}=2) \mathrm{FL}$ intervals were captured most frequently, each representing $18.2 \%$ of the total catch (Figure 2). The length-weight relationship for Lake Sturgeon captured in the Burntwood River during fall 2018 was similar to those captured in other areas and is presented in Figure 3.

There were not enough baseline data on condition factor and growth available from this area to make cohort-specific statistical pre- and post-construction comparisons of mean condition factor-at-age or fork length-at-age (Figures 4 and 5).

### 4.1.1.3 ReCAPTURES

One previously tagged juvenile Lake Sturgeon was captured in the Burntwood River (Table 8; Appendix A4-1). The fish was originally caught in 2017 and was recaptured 0.41 km downstream of its original tagging location.

### 4.1.2 Split LaKE

Twelve fish species $(\mathrm{n}=187)$ were captured at 21 sites between September 9 and 20, 2018 (Tables 2 and 3; Map 3). There were 52 juvenile and five adult Lake Sturgeon caught in 606.9 gillnet hours, producing an overall CPUE of $2.25 \mathrm{LKST} / 100 \mathrm{~m}$ net/24 h (Table 4). Lake Sturgeon were the most abundant species captured ( $n=57 ; 30.5 \%$; Table 3). Four juvenile (7.7\%) and no adult mortalities occurred during sampling. It is unclear why the number of mortalities was slightly high. Gillnet site data as well as biological and tagging information for all Lake Sturgeon captured in Zone SPL-A (Map 3) are provided in Appendices A1-1 and A2-1. Annual CPUE values for the Split Lake catch since 2015 are presented in Table 5.

### 4.1.2.1 Year-Class Strength

Ageing structures were collected from all 52 of the juvenile Lake Sturgeon caught in Split Lake. Three-reader agreement for age assignment was $96 \%(n=50)$. Aged Lake Sturgeon from Split Lake ranged from 1 to 15 years old and represented the 2003, 2005-2008, and 2010-2017 cohorts (Figure 6). No YOY (i.e., 2018 cohort) fish were captured. The 2013 cohort was the most abundant year-class, accounting for $50 \%(n=26)$ of the catch. Lake Sturgeon from the 2011 cohort were also caught relatively frequently ( $n=9 ; 17 \%$; Figure 6). Cohort frequencies for all juvenile Lake Sturgeon captured in Split Lake from 2015 to 2018 are presented in Table 6.

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Every cohort since 2003, with the exception of the 2004 and 2018 cohorts, has been present in the catch.

### 4.1.2.2 Growth and Condition

Captured Lake Sturgeon had a:

- Mean FL of $583 \mathrm{~mm}(\mathrm{n}=57$; StDev = 151 mm ; range 230-996 mm);
- Mean weight of $1,816 \mathrm{~g}(\mathrm{n}=57$; StDev $=1,428 \mathrm{~g}$; range $25-7,350 \mathrm{~g})$; and
- Mean condition factor of $0.75(\mathrm{n}=57$; StDev $=0.14$; range $0.21-1.46)$ (Table 7).

Lake Sturgeon from the 500-549 mm FL interval ( $\mathrm{n}=16$; 28.1\%) were the most frequently captured size-class (Figure 2).

There were not enough baseline data available on condition factor and growth from this area to make cohort-specific statistical pre- and post-construction comparisons of mean condition factor-at-age or fork length-at-age (Figures 4 and 5).

### 4.1.2.3 ReCAPTURES

Three previously tagged wild Lake Sturgeon were recaptured in Split Lake during sampling in 2018 (Table 8; Appendix A4-1). All three fish were originally tagged in Split Lake and were recaptured between 3.1 and 3.8 km downstream of their original capture location.

One hatchery-reared Lake Sturgeon released in the Burntwood River was captured (Table 9; Appendix A4-2):

- PIT \#900043000102957 was released as a one-year-old at stocking site 2 (Map 4) on October 2, 2014, and was captured in Zone SPL-A on September 20, 2018, 32.3 km downstream from where it was released (Zone BWR-B). It has not been captured previously during juvenile sampling. The fish grew 255 mm in length and 995 g in weight since release (Appendix A4-2).


### 4.2 Future Keeyask reservoir

Ten species ( $\mathrm{n}=401$ ) were captured at 50 sites between September 9 and 19, 2018 (Tables 2 and 3; Map 5). Water temperature decreased from $12.0^{\circ} \mathrm{C}$ to $9.0^{\circ} \mathrm{C}$ during sampling (Appendix A1-2). Lake Sturgeon ( $\mathrm{n}=150 ; 37.4 \%$ ) were the second most abundant species captured behind Longnose Sucker ( $n=158 ; 39.4 \%$; Table 3). Gillnet site data as well as biological and tagging information for all Lake Sturgeon captured are provided in Appendices A1-2 and A2-2.
In total, 143 juvenile and seven adult Lake Sturgeon were captured in 1376.8 gillnet hours, producing an overall CPUE of 2.61 LKST/100 m net/24 h (Table 4). Three juvenile (2.1\%) and no adult mortalities occurred during sampling. Gill nets were set throughout Gull Lake (i.e., in

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zones GL-A, GL-B, and GL-C), as well as the first zone upstream of Gull Lake (i.e., BR-D) (Map 5). CPUE values by zone, from upstream to downstream, were as follows:

- 0.89 LKST/ $100 \mathrm{~m} / 24 \mathrm{~h}$ in Zone BR-D ( $\mathrm{n}=4$ sites);
- $1.30 \mathrm{LKST} / 100 \mathrm{~m} / 24 \mathrm{~h}$ in Zone GL-A ( $\mathrm{n}=11$ sites);
- 3.00 LKST/100 m/24 h in Zone GL-B ( $\mathrm{n}=18$ sites); and
- 3.65 LKST/ $100 \mathrm{~m} / 24 \mathrm{~h}$ in Zone GL-C ( $\mathrm{n}=17$ sites; Table 4 ).

Total CPUE values recorded in the future Keeyask reservoir since 2007 are presented in Table 5. CPUE values in 2018 were slightly lower than 2017 but were within the range recorded in other years.

Of the 150 Lake Sturgeon captured, 17 were known hatchery-reared fish (i.e., stocked as age-1 and marked with PIT tags; discussed in further detail in Section 4.2.5). Five fish could not be accurately identified as hatchery-reared or wild fish as they had an ageing structure typical of a hatchery-reared fish but no PIT tag (not included in wild or hatchery calculations; described in Section 3.4). These fish CPUE values for wild and hatchery-reared Lake Sturgeon were as follows:

- 2.23 LKST/100 m/24 h ( $\mathrm{n}=128$ ) for wild Lake Sturgeon; and
- $0.30 \mathrm{LKST} / 100 \mathrm{~m} / 24 \mathrm{~h}(\mathrm{n}=17)$ for hatchery-reared Lake Sturgeon (Table 10).


### 4.2.1 Year-Class Strength

Ageing structures were collected from 140 of the 143 juvenile Lake Sturgeon. Ageing structures were not collected from two juveniles, that had ageing structures previously removed from both pectoral fins (described in Section 3.2), and one YOY (2018 cohort). Three-reader agreement for age assignment was $95 \%(n=133)$. The modal age was used for 7 fish that were aged one year lower or higher by a single reader (age $4[n=1]$, age $5[n=1]$, age $8[n=1]$, age $9[n=1]$ and age $10[\mathrm{n}=3]$ ). Lake Sturgeon ranged from 0 to 12 years (2006-2018 cohorts; Figure 7).

Of the 140 aged Lake Sturgeon, 118 were considered wild fish (did not exhibit weak annuli). An additional two recaptured fish did not have an ageing structure taken in 2018 but were aged when they were previously captured. These two fish were older than the age of any stocked fish (i.e., age-7 and age-10) and were therefore considered wild fish. An additional YOY captured was considered wild, as it was captured prior to the release of fingerlings into the future Keeyask reservoir in fall, 2018. Overall, 121 fish were classified as wild. An additional 17 fish were classified as known-hatchery (as they had PIT tags) and five were classified as suspected hatchery fish (exhibited weak annuli but no PIT tag).

Of the five suspected hatchery fish (i.e., not PIT tagged), two fish identified as being from the 2014 cohort, two fish from the 2015 cohort, and one fish from the 2016 cohort exhibited weak/absent first annuli, characteristic of hatchery reared-fish. If these were hatchery-reared
fish that shed their PIT tags, the fish identified as belonging to the 2015 cohort would likely belong to the 2016 cohort because, as discussed in Section 3.4, hatchery fish are sometimes aged one year above their true age (Figure 7). Based on when stocking has occurred, the two suspected hatchery fish identified as being from the 2014 cohort and the one fish from the 2016 cohort were likely aged correctly.

Of the 121 wild and 17 known hatchery-reared aged Lake Sturgeon, the 2016 cohort was captured most frequently ( $n=43 ; 31.2 \%$ ). The 2008 cohort was also relatively abundant in the catch, accounting for $23.9 \%(n=33)$; however, this is likely an underestimate of the cohort. Individuals from the 2008 cohort are now 10 years old, and some are likely > 800 mm FL. Ageing structures are not collected from individuals $>800 \mathrm{~mm}$ FL, thus more may have been present in the adult catch but were not aged. Each cohort since 2006 was present in the catch. One YOY fish (i.e., 2018 cohort) was captured in Zone GL-C. Known hatchery-reared fish accounted for $61.5 \%$ and $20.9 \%$ of the 2014 and 2016 cohorts, respectively (Figure 7).

Wild fish from all cohorts since 1998, with the exception of the 2002 cohort, have been represented in the catch since studies began (Table 6).

### 4.2.2 Population Estimate

The 2018 population estimate for the future Keeyask reservoir was 4,133 wild juvenile Lake Sturgeon ( $95 \% \mathrm{Cl}$ : $2,955-5,780$ ) (Figure 8; Appendix A5-1). This was above the $95 \%$ confidence limits of the 2012 and 2017 estimates, but within the $95 \%$ confidence limits of 2010 and 2013-2016. The estimated annual survival rate was $77 \%$.

Annual survival of the 423 hatchery-reared Lake Sturgeon stocked into the future Keeyask reservoir in 2015 (i.e., the 2014 cohort) was estimated at $95 \%$ (Appendix A5-3). Based on this estimate, 346 of these fish remained in the future Keeyask reservoir in 2018.

### 4.2.3 GROWTH AND CONDITION

In 2018, 128 wild (including adult size fish) and 17 known hatchery-reared Lake Sturgeon were captured. Length-weight relationships for hatchery-reared and wild Lake Sturgeon are presented in Figure 3.

Wild Lake Sturgeon had a:

- Mean FL of $524 \mathrm{~mm}(\mathrm{n}=128$; StDev = 206 mm ; range 87-1,031 mm);
- Mean weight of $1,565 \mathrm{~g}(\mathrm{n}=127$; StDev $=1,634 \mathrm{~g}$; range $50-8,500 \mathrm{~g})$; and
- Mean condition factor of $0.73(n=127 ;$ StDev $=0.13$; range $0.32-1.30)$ (Table 7).

Wild Lake Sturgeon in the 300-349 mm FL interval were captured most frequently, representing $19.5 \%(n=25)$ of the wild catch (Figure 11). Fish measuring 700-749 mm were also frequently captured representing $13.3 \%(n=17)$ of the wild catch (Figure 11).
Mean condition factor-at-age was not significantly different at any age for Lake Sturgeon captured during construction (i.e., 2014-2018) and baseline (i.e., 2008-2012) studies (MannWhitney $U$ test and $t$-test, $p>0.05$; Figure 9). Mean FL-at-age was significantly lower for age-7 and age-2 Lake Sturgeon but was higher for age-1 fish captured during construction than during baseline studies (Mann-Whitney $U$ test, $t$-test, $p<0.05$ ) (Figure 10). A comparison of the von Bertalanffy growth curves indicated that fish captured during baseline and construction grew at similar rates (Figure 13).

Hatchery-reared Lake Sturgeon had a:

- Mean FL of $396 \mathrm{~mm}(\mathrm{n}=17$; StDev $=57 \mathrm{~mm}$; range 255-479 mm);
- Mean weight of $394 \mathrm{~g} \mathrm{( } \mathrm{n}=17$; StDev = 148 g ; range $100-700 \mathrm{~g}$ ); and
- Mean condition factor of $0.60(\mathrm{n}=17$; StDev $=0.05$; range $0.53-0.72)$ (Table 7).

Hatchery-reared Lake Sturgeon in the 400-449 mm FL interval were captured most frequently, representing $41.2 \%$ of the hatchery catch $(\mathrm{n}=7)$ (Figure 11).
Mean FL-at-age for all four age classes (ages 1-4) were significantly higher for hatchery-reared Lake Sturgeon when compared to wild fish (Mann-Whitney $U$ test, $p<0.05$; Figure 12). Mean weight-at-age was significantly higher in age-1 and age-2 hatchery fish and significantly lower in age-4 hatchery fish when compared to wild fish (Mann-Whitney $U$ test, $p<0.05$; Figure 12). Mean condition factor-at-age was significantly lower for age-1, age-3, and age-4 hatcheryreared fish when compared to wild Lake Sturgeon (Mann-Whitney U test, $p<0.05$; Figure 12).

A comparison of von Bertalanffy growth curves between hatchery and wild fish showed that young hatchery fish are longer and heavier than wild fish of the same cohort. However, growth of hatchery fish appears to slow around age-4 (Figure 14).

### 4.2.4 ReCAPTURES

Eighteen Lake Sturgeon tagged in previous years were captured (Table 8; Appendix A4-1). All fish were originally tagged in the future Keeyask reservoir: one in 2006, one in 2010, one in 2011, one in 2013, seven in 2014, three in 2016, and four in 2017.

Recaptured fish moved varying distances from their original capture locations:

- Five moved less than 1.0 km;
- Twelve were recaptured within 1.0-9.2 km; and
- One was recaptured 21.2 km downstream of its original capture location in Zone BR-D.


### 4.2.5 Hatchery Captures

Seventeen known hatchery fish (i.e., those PIT tagged and stocked as age-1) were caught in 2018, representing $11 \%$ of the total catch (Table 9; Appendix A4-2). None of the hatcheryreared fish had been captured during previous sampling. An age breakdown of all the hatcheryreared fish captured between 2014 and 2018 is presented in Table 11.

Of the 17 hatchery fish:

- Eight were stocked in Gull Lake in 2015 (Table 9):
- Three were stocked at Site 1 on June (Zone GL-C; Map 6) and were caught between 2.72 and 2.80 km downstream.
- One was stocked at Site 2 on June 22 (Zone GLC; Map 6) and was captured 0.39 km downstream.
- Four were stocked at Site 6 on September 15 (Zone GL-B; Map 6). Three were captured less than 0.5 km from their release location and one was captured 3.92 km downstream.
- Eight were stocked on June 8, 2017 in Zone GL-A (Site 1; Map 6):
- Three were caught in Zone GL-A less than 2.7 km from their release location.
- Four were caught in Zone GL-B between 8.5 and 9.2 km downstream.
- One was caught in Zone GL-C, approximately 12 km downstream of its original release location.
- One was stocked in the Burntwood River on June 7, 2018 (Site 2, Map 4). The fish was captured on September 19, 2018, in Zone GL-C, 127.8 km downstream of its release location. The fish grew 62 mm in length and 53 g in weight since release.


### 4.3 Stephens Lake

Nine fish species ( $n=430$ ) were captured at 49 gillnetting sites in upper Stephens Lake between September 9 and 21, 2018 (Tables 2 and 3; Map 7). Lake Sturgeon ( $n=74$; 17.2\%) were the third most abundant species captured, after Longnose and White suckers (Table 3). Gillnet site data, as well as biological and tagging information for all Lake Sturgeon captured in Stephens Lake, are provided in Appendices A1-3 and A2-3.

In total, 71 juvenile and 3 adult Lake Sturgeon were captured in 1,599.0 gillnet hours, producing an overall CPUE of $1.11 \mathrm{LKST} / 100 \mathrm{~m}$ net/24 h (Table 4). No mortalities occurred during sampling. Gill nets were set in both zones located within the upper 10 km of Stephens Lake (Map 7). CPUE values by zone were as follows:

- 0.79 LKST/100 m/24 h in Zone STL-A ( $\mathrm{n}=17$ sites); and
- 1.29 LKST/100 m/24 h in Zone STL-B ( $\mathrm{n}=32$ sites; Table 4).

CPUE values for the Stephens Lake sturgeon catches since 2007 are presented in Table 5. CPUE in Stephens Lake was lower than in 2017 but was similar to other construction monitoring years (2014-2016).

Of the 74 Lake Sturgeon, 17 were known hatchery fish (i.e., stocked at age-1 and marked with PIT tags; discussed in further detail in Section 4.3.5). Five fish could not be accurately identified as hatchery or wild fish as they had an ageing structure typical of a hatchery-reared fish but no PIT tag (not included in wild or hatchery calculations; described in Section 3.4). CPUE values for wild Lake Sturgeon and hatchery-reared Lake Sturgeon were as follows:

- 0.78 LKST/ $100 \mathrm{~m} / 24 \mathrm{~h}(\mathrm{n}=52)$ for wild Lake Sturgeon; and
- 0.26 LKST/100 m/24 h $(\mathrm{n}=17)$ for hatchery-reared Lake Sturgeon (Table 10).


### 4.3.1 Year-Class Strength

Ageing structures were collected from 66 of the 71 juvenile Lake Sturgeon captured in Stephens Lake. Ageing structures were not taken from three fish that had ageing structures previously removed from both pectoral fins, one fish that had deformed fins, and one fish that was identified as a hatchery-reared fish in the field and therefore its age was known. Three-reader agreement for age assignment was $93.9 \%(n=62)$. The modal age was used for 4 fish aged one year lower or higher by a single reader (age $2[n=1], 4[n=1], 5[n=1], 10[n=1]$ ). Aged juvenile Lake Sturgeon ranged from 1-10 years. Of the 70 aged Lake Sturgeon (including 66 aged in 2018, three aged in previous years, and one known-age hatchery fish), 48 were considered wild fish (did not exhibit weak annuli), 5 were classified as suspected hatchery fish (exhibit weak annuli but no PIT tag), and 17 were known hatchery fish (with PIT tags).

Five fish without hatchery implanted PIT tags exhibited weak/absent first annuli, characteristic of hatchery-reared fish. Of the five fish, four were aged as three-year-olds and one was aged as a five-year old. It is likely these fish were stocked as yearlings and shed their PIT tags. As a result, the four three-year-old fish have likely been over-aged and are believed to belong to the 2016 cohort (age-2 fish) and not the 2015 cohort (Figure 15). Similarly, the one five-year old fish likely belongs to the 2014 cohort (age-4) and not the 2013 cohort (age-5).

The 2015 cohort (age-3) was the most frequent in the catch accounting for $28.6 \%(n=20)$ of the aged fish (Figure 15). The 2014, 2016, and 2013 cohorts (ages 4, 2 and 5) were the next most abundant age-classes, representing $18.6 \%(n=13), 15.7 \%(n=11)$, and $12.9 \%(n=9)$ of the catch, respectively (Figure 15). Known hatchery-released fish accounted for the majority of fish caught from the 2014 ( $76.9 \%$; $n=10$ ) and 2016 ( $63.6 \%$; $n=7$ ) cohorts (Figure 15). In 2018, all cohorts from 2011-2017 were present in the catch. However, all cohorts between 2000 and 2017 have been represented in the catch since studies began (Table 6).

### 4.3.2 POPULATION Estimate

The 2018 population estimate for Stephens Lake was 1,101 juvenile Lake Sturgeon ( $95 \% \mathrm{CI}$ : 749-1,620) (Figure 16; Appendix A5-2). This was above the 95\% confidence limits of the 2014 and 2016 estimates, but within the $95 \%$ confidence limits of 2010-2013 and 2017. The estimated annual survival rate was $88 \%$.

Survival of the 418 hatchery-reared Lake Sturgeon stocked into Stephens Lake in 2015 (i.e., the 2014 cohort) was estimated at $86 \%$ (Appendix A5-3). Based on this estimate, 233 of these fish remained in Stephens Lake in 2018.

### 4.3.3 GROWTH AND CONDITION

In 2018, 52 wild (including adult size fish) and 17 known hatchery-reared Lake Sturgeon were caught. Length-weight relationships for hatchery-reared and wild Lake Sturgeon are presented in Figure 3.

Wild Lake Sturgeon had a:

- Mean FL of $488 \mathrm{~mm}(\mathrm{n}=52 ; \operatorname{StDev}=158 \mathrm{~mm}$; range 222-837 mm);
- Mean weight of $1,172 \mathrm{~g}(\mathrm{n}=52$; StDev $=1,254 \mathrm{~g}$; range $50-4,925 \mathrm{~g})$; and
- Mean condition factor of $0.72(\mathrm{n}=52$; StDev $=0.10$; range $0.46-0.90)$ (Table 7).

Wild Lake Sturgeon in the 400-449 mm FL interval were captured most frequently accounting for $28.8 \%(n=15)$ of the wild catch. The $500-549 \mathrm{~mm}$ and $300-349 \mathrm{~mm}$ FL intervals were also caught frequently and accounted for $15.4 \%(\mathrm{n}=8)$ and $13.5 \%(\mathrm{n}=7)$ of the wild catch, respectively (Figure 11).

Mean condition factor-at-age was significantly lower for age-3 Lake Sturgeon captured during construction (i.e., 2014-2018) than during baseline (i.e., 2009-2012) studies (Mann-Whitney U test, $p<0.05$; Figure 17). Mean FL-at-age was significantly higher for age-3 fish and significantly lower for age-4 fish caught during construction when compared to baseline studies; however, due to small sample sizes, only age 2, 3, and 4 fish could be compared (MannWhitney $U$ test; $p<0.05$; Figure 18).

Hatchery-reared Lake Sturgeon had a:

- Mean FL of $432 \mathrm{~mm}(\mathrm{n}=17$; StDev $=64 \mathrm{~mm}$; range $346-503 \mathrm{~mm})$;
- Mean weight of $596 \mathrm{~g}(\mathrm{n}=17$; StDev = 239 g ; range 275-900 g); and
- Mean condition factor of $0.69(n=17 ; S t D e v=0.08$; range $0.48-0.85)$ (Table 7).

Hatchery-reared Lake Sturgeon in the 400-449 mm FL interval were captured most frequently ( $47.1 \% ; \mathrm{n}=8$ ) (Figure 11). Fish in the $350-399 \mathrm{~mm}$ FL interval were also frequently caught (29.4\%; $n=5$ ).

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Mean FL-at-age was significantly higher for age-1, age-2 and age-3 hatchery-reared Lake Sturgeon than for wild fish (Mann-Whitney U test, $p<0.05$; Figure 19). Mean weight-at-age was significantly higher for age-1 hatchery fish but was significantly lower in age-3 hatchery fish when compared to wild fish (Mann-Whitney $U$ test, $p<0.05$; Figure 19). Mean condition factor-at-age was significantly lower for age-3 and age-4 hatchery-reared Lake Sturgeon than for wild fish (Mann-Whitney U test, $p<0.05$; Figure 17).

A comparison of von Bertalanffy growth curves between hatchery and wild fish showed that young hatchery fish are longer and heavier than wild fish of the same cohort. However, growth of hatchery fish appears to slow around age-4 (Figure 20)

### 4.3.4 ReCAPTURES

Eleven Lake Sturgeon tagged in a previous year were recaptured in Stephens Lake (Table 8; Appendix A4-1). Of these: one was tagged in 2011, one in 2014, two in 2015, three in 2016, three in 2017, and one in spring of 2018.

Recaptured fish moved varying distances from their original capture locations:

- Ten fish were originally captured in Stephens Lake. All were recaptured close to their original capture locations.
- Five were recaptured within 1.1 km .
- Five were recaptured between 2.3 and 3.1 km .
- One fish (Floy \#96513) was tagged in the future Keeyask reservoir (Zone GL-C) 15.0 km upstream as an age-2 in 2014 and has been caught multiple times since. It was recaptured in 2015 in the future Keeyask reservoir (GL-C), in 2017 in Stephens Lake (STL-A), and in 2018 in Stephens Lake (STL-B). It increased in size by 188 mm and 775 g since its initial capture in 2014. This fish represents one of only three wild fish from the future Keeyask reservoir captured downstream in Stephens Lake (Burnett et al. 2018).


### 4.3.5 Hatchery Captures

A total of 17 hatchery-reared Lake Sturgeon released as one-year-olds were captured in Stephens Lake in 2018, representing 23\% of the total catch (Table 9; Appendix A4-2). None of the hatchery-reared fish had been captured during previous sampling. An age breakdown of all the hatchery-reared fish captured between 2014 and 2018 is presented in Table 11.

Of the 17 hatchery captures:

- Fourteen were stocked in Stephens Lake:
- Two were stocked on June 22, 2015, at Site 3 (Zone STL-B; Map 8). One was captured 1.8 km downstream and one 2.3 km upstream of the original release location.
- Five were stocked on September 14, 2015.
- Three were stocked in Zone STL-A (Site 5; Map 8) of which one was caught 0.7 km upstream and two were caught 0.7 km downstream.
- Two were stocked in Zone STL-B (Site 4; Map 8) and were captured 0.9 km and 3.3 km upstream.
- Seven were stocked at the Keeyask downstream boat launch (Site 1; Map 8) on June 15, 2017. All seven were captured in Zone STL-B, 1.7 to 4.0 km downstream.
- Three were stocked in the future Keeyask reservoir:
- One was stocked on June 22, 2015, in Zone GL-B (Map 6) and was captured 15.0 km downstream.
- Two were stocked on September 16, 2015.
- One was released at Site 6 (GL-B; Map 6) and was captured 13.1 km downstream.
- One was released at Site 7 (GL-C; Map 6) and was captured 11.2 km downstream.


### 5.0 DISCUSSION

### 5.1 Juvenile Abundance

In 2018, an increase in gillnetting effort resulted in an increase in the number of sturgeon captured in Split Lake compared to previous years. It is not known whether these fish were hatched in the Burntwood or Nelson rivers, although fish captured along the south shore of Split Lake are in the main flow of the Nelson River and are more likely to have originated there. The CPUE and number of juvenile Lake Sturgeon captured in the Burntwood River in 2018 was lower than in previous years; however, catches have been variable with no clear trend since the program started. Inter-annual variation in flow also has a substantial effect on where nets can be set; in 2018, 19 gill nets were set (compared to 24 in 2017) due to a reduced number of suitable sites and the presence of debris.
In both the future Keeyask reservoir and Stephens Lake CPUE was slightly lower in 2018 (2.61 and $1.11 \mathrm{LKST} / 100 \mathrm{~m} / 24 \mathrm{~h}$ ) compared to 2017 ( 2.74 and 1.98 ). Environmental factors may have influenced juvenile capture, as Nelson River flows were lower ( $\sim 2,800 \mathrm{~m}^{3} / \mathrm{s}$ ) during the 2018 study period than in the previous two years (> $4,000 \mathrm{~m}^{3} / \mathrm{s}$ ). Although flow fluctuates on a seasonal and yearly basis, the juvenile catch in both areas was higher in years of high flow (i.e., 2016 and 2017). It is possible that high flows reduced suitable habitat areas, concentrating juvenile Lake Sturgeon into smaller areas and making them easier to catch. In Stephens Lake, juvenile Lake Sturgeon captures may also have been impacted by changes to flows caused by the Keeyask Spillway commissioning in August 2018. Juvenile Lake Sturgeon tagged with acoustic transmitters spent more time in the area closer to Gull Rapids after the spillway was commissioned in August than in any previous study year (Lacho and Hrenchuk 2019). Similarly, in the current study, fewer juvenile Lake Sturgeon were captured in sites fished in previous years and more were captured closer to the construction site. However, this observation was also coincident with declining flows on the Nelson River (Section 2.2).

Mark recapture data was used to generate a population estimate of wild juvenile Lake Sturgeon in the future Keeyask reservoir and Stephens Lake for the first time in 2018. The 2018 estimate was 4,133 individuals ( $95 \% \mathrm{Cl}: 2,955-5,780$ ) for the future Keeyask reservoir and 1,101 individuals ( $95 \% \mathrm{Cl}$ : 749-1,620) for Stephens Lake. Survival was slightly lower (77\%) in the future Keeyask reservoir than in Stephens Lake (88\%). As this was the first year that the sample sizes were large enough to calculate an estimate, these numbers will likely change as data are added to the model and estimates are refined.

### 5.2 ReCRUITMENT

Benchmarks set in the Keeyask AEMP for juvenile Lake Sturgeon year-class strength relied on comparing CPUE to the Sea Falls to Sugar Falls reach of the Nelson River. When the AEMP was written, there was little information on juvenile Lake Sturgeon in the lower Nelson River, therefore benchmarks were chosen to compare to Sea Falls, where more was known. In recent years, successful stocking in the Sea Falls reach has led to a dramatic increase in numbers of juvenile Lake Sturgeon, thus, the two areas are no longer comparable. The early warning trigger for juvenile Lake Sturgeon year-class-strength has been revised to be the absence of wild juvenile Lake Sturgeon $\leq 3$ years of age in the catch in a single year. The absence of wild fish $\leq 3$ years of age in the catch for two consecutive monitoring years will be considered the ecologically significant benchmark. These new thresholds were selected based on the observation that sturgeon less than three years old have been observed in the Upper Split Lake Area, the future Keeyask reservoir and Stephens Lake in all recent years. In 2018, 85 wild Lake Sturgeon aged between 0-3 were caught: eight in the Upper Split Lake Area, 50 in the future Keeyask reservoir, and 27 in Stephens Lake.

Based on similarities in cohort frequency data as well as genetic evidence (Gosselin et al. 2016), Henderson et al. (2015) hypothesized recruitment to the Stephens Lake population was mainly due to spawning upstream of Gull Rapids (i.e., in the future Keeyask reservoir). However, more recent results indicate the cohort frequency distribution in Stephens Lake differs from that in the future Keeyask reservoir. The 2015 cohort continues to be strong in Stephens Lake, representing $27 \%$ of the wild catch in 2017 and $42 \%$ of the wild catch in 2018. The same cohort made up only $7 \%$ of the wild catch in the future Keeyask reservoir in both 2017 and 2018. Burnett et al. (2018) hypothesized that if these fish had been spawned upstream of Gull Rapids and drifted downstream, a larger proportion would have remained upstream. It is likely that spawning at Gull Rapids contributed to the 2015 cohort in Stephens Lake. This observation is consistent with the recent increase in the number of mature males captured during spring adult surveys below Gull Rapids (Legge et al. 2017; Holm and Hrenchuk 2019).
Hatchery-reared Lake Sturgeon released as yearlings are also contributing to a large portion of the 2014 and 2016 cohorts in both the future Keeyask reservoir and Stephens Lake. These hatchery fish made up $62 \%$ of the 2014 cohort and $21 \%$ of the 2016 cohort in the future Keeyask reservoir, and $77 \%$ of the 2014 cohort and $64 \%$ of the 2016 cohort in Stephens Lake. As noted earlier, hatchery raised sturgeon released as larvae or fingerlings may also contribute to the total cohorts, but these cannot be distinguished from the wild fish without additional testing.

### 5.3 Hatchery Fish

The stocking program began in 2014, with regular stocking events in the Burntwood River, the future Keeyask reservoir and Stephens Lake since then. In the spring of 2018, 739 age-1 (PIT
tagged) Lake Sturgeon were released in the Burntwood River bringing the total number of stocked age-1 fish in this area to 1,357 at the time of the current study. A total of 886 and 1,138 age-1 Lake Sturgeon were stocked in the future Keeyask reservoir and Stephens Lake, respectively, prior to 2018. Hatchery-reared (i.e., age-1 fish marked with PIT tags) Lake Sturgeon continued to make up a large proportion of the catch in both the future Keeyask reservoir (11\%) and Stephens Lake (23\%) in 2018.

Despite large number of hatchery captures in the future Keeyask reservoir and Stephens Lake, few hatchery-reared fish have been captured in the Burntwood River. In total, six hatcheryreared fish ( $0.4 \%$ of stocked fish) have been captured in the Upper Split Lake Area: one in 2014, one in 2016, three in 2017, and one in 2018. Only four of these fish were captured in the Burntwood River itself. Four yearlings stocked in the Burntwood River have been captured in the future Keeyask reservoir, which suggests the majority of yearlings stocked in the Burntwood River are not remaining there, but are likely moving downstream. Additional sites were added to Split Lake in 2018 in an attempt to capture hatchery fish, but only a single hatchery-reared Lake Sturgeon was captured. This fish was captured in the southern portion of Split Lake approximately 32 km downstream of where it was stocked (GN-38; Map 3; Appendix A1-1).

Some fish stocked in the future Keeyask reservoir have moved downstream into Stephens Lake. In 2017, $65 \%(\mathrm{n}=20)$ of the captured yearlings originally stocked in the future Keeyask reservoir were captured in the future Keeyask reservoir, and $35 \%$ ( $n=11$ ) were captured in Stephens Lake. In 2018, $84 \%(\mathrm{n}=16)$ of the captured yearlings originally stocked in the future Keeyask reservoir were captured in the future Keeyask reservoir and $16 \%(n=3)$ were captured in Stephens Lake. This pattern of downstream movement is not found in recaptured wild fish, as only three wild juvenile Lake Sturgeon tagged in the future Keeyask reservoir have ever been captured in Stephens Lake. Additionally, no juvenile Lake Sturgeon tagged with acoustic transmitters have moved downstream through Gull Rapids during five years of study (Lacho and Hrenchuk 2019). Burnett et al. (2018) hypothesized that observed downstream movements of hatchery fish in 2017 may have been at least partially due to extreme flows. Future monitoring will determine whether future cohorts stocked in less extreme flow events continue to move downstream.

Similar to 2017, several Lake Sturgeon in $2018(\mathrm{n}=10)$ had ageing structures characteristic of hatchery-reared fish (i.e., weak or missing first annuli) but did not have hatchery implanted PIT tags. Although the ages of PIT tagged hatchery fish are known, obtaining the correct age from pectoral fin rays has proven difficult. Of the 34 known-age hatchery-reared fish captured in the future Keeyask reservoir and Stephens Lake in 2018, all (three-reader agreement at 88.2\%) were over-aged by one year due to the presence of a weak first annuli in combination with the use of a correction for the missing first annulus (see Section 3.4). Ageing structure analysis suggests PIT tag loss is occurring due to the high number of suspected hatchery fish in the catch. In the future Keeyask reservoir, suspected hatchery fish accounted for 13\% of the 2014 year-class and 18\% of the 2015 year-class (which is likely the 2016 year class). In Stephens Lake, suspected hatchery fish accounted for $10 \%$ of the 2013 year-class (likely 2014 year class) and $17 \%$ of the 2015 year-class (likely 2016 year-class).

### 5.4 Key Questions

The AEMP identified key questions for juvenile Lake Sturgeon monitoring, four of which are relevant to the construction period and are addressed in the discussion below.

Does recruitment of wild sturgeon occur upstream and/or downstream of the GS during construction?

In 2018, one wild YOY sturgeon was caught in the future Keeyask reservoir and none were caught in Stephens Lake. While the presence of one YOY sturgeon in the future Keeyask reservoir indicates successful spawning occurred upstream of Gull Rapids, the absence of YOY in Stephens Lake does not necessarily mean recruitment was unsuccessful downstream of Gull Rapids. YOY are often underrepresented in the catch due to their size and the gear used does not adequately target YOY fish. All other cohorts spawned since construction started (i.e., 2015-2017) have been present in the catch in both areas.

Is there a biologically meaningful (and statistically significant) change in condition factor and growth of juvenile sturgeon during construction?

A comparison of juvenile Lake Sturgeon condition factor between baseline and construction monitoring studies found significant differences in several age classes; however, as in 2017, there is not a consistent trend by location or age-class. In the Upper Split Lake Area, no statistical comparisons could be made due to a lack of baseline data. A qualitative examination of the data suggests no obvious differences in condition factor- or FL-at-age.

Although no clear trends in FL-at-age and condition factor-at-age could be seen among sampling years, differences appear to be consistent between reporting years. However, these differences may be due to small baseline sample sizes and not be biologically significant. For example, age-7 Lake Sturgeon captured in the future Keeyask reservoir have been significantly shorter during construction than baseline since 2015. However, only 10 age-7 fish were captured during baseline, compared to a much larger sample ( $\mathrm{n}=82$ ) during construction.

A comparison of growth curves of wild fish captured in the future Keeyask reservoir during baseline and construction was conducted for the first time in 2018. Growth during both time periods was similar. Too few juveniles were collected in Stephens Lake prior to construction to support a pre/post analysis.

Two questions related to the stocking program are addressed below:

- What is the survival rate of stocked sturgeon?; and
- What is the proportion of hatchery-reared to wild recruits within a cohort (i.e., how successful is the stocking program)?

As noted previously, only sturgeon stocked as yearlings can be distinguished from wild fish and the following discussion considers only fish stocked as yearlings. In 2018 the annual survival rate of stocked sturgeon was estimated as $95 \%$ in the future Keeyask reservoir and $86 \%$ in

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Stephens Lake. As the number of recaptures increases over subsequent study years, the survival rate will be adjusted to account for newly captured fish.

The proportion of hatchery-reared Lake Sturgeon captured in both the future Keeyask reservoir and Stephens Lake in 2018 remained high. In the future Keeyask reservoir, hatchery-reared Lake Sturgeon made up $2 \%$ of the catch in 2015, $7 \%$ in 2016, 12\% in 2017, and $11 \%$ in 2018. Hatchery-reared fish tend to make up a larger proportion of the catch in Stephens Lake: 7\% in $2015,8 \%$ in 2016, $34 \%$ in 2017, and $23 \%$ in 2018. In Stephens Lake, the 2014 and 2016 cohorts were dominated by hatchery-reared fish, $77 \%$ and $64 \%$, respectively.

The recapture of stocked fish with PIT tags indicates hatchery-reared sturgeon are growing and surviving in the wild. Statistical analyses of FL-, weight-, and condition factor-at-age as well as von Bertalanffy growth curve analysis suggest that hatchery-reared Lake Sturgeon in both areas may be growing differently than wild fish. Hatchery fish were both longer and heavier at release than wild fish of the same age, and this "headstart" appears to persist until age-4 (when the accelerated growth rate tapers off). However, to date, the oldest hatchery-reared Lake Sturgeon stocked into both the future Keeyask reservoir and Stephens Lake are 4 years old (2014 cohort). Continued monitoring of these fish will determine if these differences persist as the fish grow.

### 5.5 Next Steps

The juvenile Lake Sturgeon population monitoring program will be repeated in 2019. Sampling locations in the Burntwood River will remain similar to previous years, but will reflect the change made in 2018 to focus sampling effort on Split Lake and not the reach of the Nelson River immediately downstream of the Kelsey GS. Increased netting effort in Split Lake (downstream of the Nelson and Burntwood rivers) will identify whether a large number of unidentified juveniles occur in the lake, and may also aid in locating juveniles stocked in the Burntwood River that have moved downstream. Focus in Split Lake will be on sites not sampled in 2018, as well as those in the vicinity of where the single hatchery fish was captured. The number of sites in the Burntwood River will be increased in order maintain similar effort as in previous sampling years. Sampling in the future Keeyask reservoir and Stephens Lake will remain similar to previous years.

Comparative analysis of juvenile Lake Sturgeon growth rate prior to and during construction was conducted for the first time in 2018 and will continue to be assessed as more data is added. Growth of hatchery fish will continue to be monitored using von Bertalanffy growth curves with growth estimates likely to become more refined as the length of time hatchery fish will have spent in the wild increases. Population estimates were also generated for the first time in 2018 and will be continually updated as more mark recapture data becomes available in subsequent study years.

### 6.0 SUMMARY AND CONCLUSIONS

- Sampling locations in the Burntwood River, the future Keeyask reservoir, and Stephens Lake remained similar to previous years; however, the number of sampling locations decreased in 2018 due to flows and the presence of debris. Sampling was not conducted in the Kelsey GS area, as effort was shifted to Split Lake in an attempt to captured stocked fish.
- Sixty eight Lake Sturgeon were captured in the Upper Split Lake Area: 11 in the Burntwood River (426.3 gillnet hours, CPUE of 0.62 Lake Sturgeon/100 m net/24 h) and 57 (52 juvenile and 5 adult) in Split Lake ( 606.9 gillnet hours, CPUE of 2.25 Lake Sturgeon/ 100 m net/ 24 h ). One of the five recaptured fish was a hatchery-reared fish released into the Burntwood River in 2014 (2013 cohort; captured in Split Lake). No YOY (2018 cohort) were caught in the Upper Split Lake Area in 2018.
- In the future Keeyask reservoir, 150 (143 juvenile and seven adult) Lake Sturgeon were captured in 1,376.8 gillnet hours for a total CPUE of 2.61 Lake Sturgeon/100 m net/24 h. Aged Lake Sturgeon $(\mathrm{n}=143)$ ranged from 0 to 12 years old with two-year-old fish (2016 cohort) the most prevalent in the catch ( $n=44 ; 30.8 \%$ ). Eighteen Lake Sturgeon tagged in previous years and 17 stocked yearlings were captured. One of the hatchery fish was released in the Burntwood River in 2014 (2013 cohort), and was caught 127.8 km downstream of its release location. Eight of the captured stocked yearlings were released into Gull Lake in 2015 (2014 cohort) and eight were released in 2017 (2016 cohort).
- In Stephens Lake, 74 ( 71 juvenile and 3 adult) Lake Sturgeon were captured in 1,599.0 gillnet hours for a total CPUE of 1.11 Lake Sturgeon/100 m net/24 h. Lake Sturgeon ages ranged from 1 to 10 with the 2015 cohort (age-3) captured most frequently ( $\mathrm{n}=24$; $34.3 \%$ ). Eleven Lake Sturgeon tagged in a previous year and 17 stocked yearlings were captured. Three of the stocked yearlings were released in the future Keeyask reservoir in 2015 (2014 cohort) and 14 were released in Stephens Lake (7 in 2015 [2014 cohort] and 7 in 2017 [ 2016 cohort]).
- Abundance estimates were calculated for wild juvenile Lake Sturgeon in the future Keeyask reservoir and Stephens Lake for the first time in 2018. The population estimate in 2018 was 4,133 wild juvenile Lake Sturgeon ( $95 \% \mathrm{CI}: 2,955-5,780$ ) for the future Keeyask reservoir and 1,101 ( $95 \% \mathrm{Cl}$ : 749-1,620) for Stephens Lake. Survival was estimated as $77 \%$ in the future Keeyask reservoir, and $88 \%$ in Stephens Lake.
- Ageing structures collected from stocked yearlings have unique banding patterns increasing the likelihood of over-ageing hatchery fish. Hatchery implanted PIT tags are crucial in determining the known ages of stocked yearlings but it appears PIT tag loss can be high. Additional analysis (genetic or fin-ray micro-chemistry) may provide a useful tool for discriminating hatchery fish from wild caught sturgeon in the case of PIT tag loss.
- The key questions, as described in the AEMP, for juvenile Lake Sturgeon population monitoring during construction of the Keeyask GS are as follows:
- Does recruitment of wild sturgeon occur upstream and/or downstream of the GS during construction?

In 2018, one wild YOY sturgeon was caught in the future Keeyask reservoir and none were caught in Stephens Lake. In both areas, wild Lake Sturgeon spawned during construction have been captured in all monitoring years (2015-2017) and wild sturgeon from these cohorts continue to captured during the 2018 monitoring, indicating that recruitment continues to occur during construction.

- Is there a biologically meaningful (and statistically significant) change in condition factor and growth of juvenile sturgeon during construction?

A comparison of mean condition factor- and FL-at-age for juvenile Lake Sturgeon collected during baseline and construction monitoring studies found significant differences in several age classes in both the future Keeyask reservoir and Stephens Lake. However, due to the lack of consistent findings in both FL-at-age and condition factor-at-age across years and waterbodies, the observed differences are not considered biologically meaningful. Growth curves for fish captured in the future Keeyask reservoir were similar during baseline and construction monitoring studies.

- What is the survival rate of stocked sturgeon? What is the proportion of hatcheryreared to wild recruits within a cohort (i.e., how successful is the stocking program)?

The survival rates of stocked sturgeon in the future Keeyask reservoir and Stephens Lake were $95 \%$ and $86 \%$, respectively. The proportion of stocked fish in the catch from the future Keeyask reservoir and Stephens Lake remained high in 2018. In Stephens Lake and, to a lesser extent, the future Keeyask reservoir, the 2014 and 2016 cohorts were dominated by hatchery fish.

Analysis of growth between hatchery and wild caught fish showed that, in both areas, young hatchery fish are longer and heavier than wild fish of the same cohort. However, growth of hatchery fish appears to slow around age-4.

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## TABLES

Table 1: $\quad$ Summary of Lake Sturgeon stocking since 2014. Numbers of stocked fish are from Klassen et al. 2019.

| Year ${ }^{\text {a }}$ | Burntwood River |  |  | Future Keeyask Reservoir ${ }^{\text {b }}$ |  |  | Stephens Lake |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Larvae | Fingerlings | Age-1 | Larvae | Fingerlings | Age-1 | Larvae | Fingerlings | Age-1 |
| 2014 | - | - | 595 | 152,926 | 4,656 | - | - | - | - |
| 2015 | - | - | - | - | - | 423 | - | - | 418 |
| 2016 | - | - | 23 | 192,167 | 780 | - | 184,134 | 799 | - |
| 2017 | 71,740 | 3,765 | - | - | - | 463 | - | - | 720 |
| 2018 | - | - | 739 | - | (933) ${ }^{\text {c }}$ | - | - | $(1,010)^{\text {c }}$ | - |
| Total | 71,740 | 3,765 | 1,357 | 345,093 | 6,369 | 886 | 184,134 | 1,809 | 1,138 |

a - Stocking year.
b - From Birthday Rapids to Gull Rapids.
c - numbers in parentheses were stocked after the 2018 juvenile survey.

Table 2: Summary of start and completion dates for juvenile Lake Sturgeon monitoring during fall, 2018, by location.

| Location | Start Date | Completion Date | \# Sites |
| :--- | :--- | :--- | :--- |
| Upper Split Lake Area |  |  |  |
| Burntwood River | 12-Sep-18 | 17-Sep-18 | 19 |
| Split Lake | $09-S e p-18$ | $20-$ Sep-18 | 21 |
| Future Keeyask Reservoir | $09-S e p-18$ | $19-$ Sep-18 | 50 |
| Stephens Lake | $09-S e p-18$ | $21-$ Sep-18 | 49 |

Table 3: $\quad$ Number ( $n$ ) and frequency of occurrence (\%), by species and sampling location, of fish captured during juvenile Lake Sturgeon monitoring, fall 2018.

| Species | Scientific Name | Upper Split Lake Area |  |  |  | Future Keeyask Reservoir |  | Stephens Lake |  | Total n | Total \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Burntwood River |  | Split Lake |  |  |  |  |  |  |  |
|  |  | n | \% | n | \% | n | \% | n | \% |  |  |
| Burbot | Lota lota | 2 | 4.1 | 13 | 7.0 | 0 | 0.0 | 49 | 11.4 | 64 | 6.0 |
| Lake Chub | Couesius plumbeus | 0 | 0.0 | 0 | 0.0 | 1 | 0.2 | 0 | 0.0 | 1 | 0.1 |
| Lake Sturgeon | Acipenser fulvescens | 11 | 22.4 | 57 | 30.5 | 150 | 37.4 | 74 | 17.2 | 292 | 27.4 |
| Lake Whitefish | Coregonus clupeaformis | 0 | 0.0 | 3 | 1.6 | 2 | 0.5 | 12 | 2.8 | 17 | 1.6 |
| Longnose Sucker | Catostomus catostomus | 24 | 49.0 | 32 | 17.1 | 158 | 39.4 | 180 | 41.9 | 394 | 36.9 |
| Northern Pike | Esox lucius | 0 | 0.0 | 1 | 0.5 | 6 | 1.5 | 0 | 0.0 | 7 | 0.7 |
| Sauger | Sander canadensis | 3 | 6.1 | 25 | 13.4 | 15 | 3.7 | 5 | 1.2 | 48 | 4.5 |
| Shorthead Redhorse | Moxostoma macrolepidotum | 0 | 0.0 | 2 | 1.1 | 9 | 2.2 | 2 | 0.5 | 13 | 1.2 |
| Spottail Shiner | Notropis hudsonius | 0 | 0.0 | 1 | 0.5 | 0 | 0.0 | 0 | 0.0 | 1 | 0.1 |
| Trout-perch | Percopsis omiscomaycus | 0 | 0.0 | 2 | 1.1 | 8 | 2.0 | 4 | 0.9 | 14 | 1.3 |
| Walleye | Sander vitreus | 2 | 4.1 | 27 | 14.4 | 5 | 1.2 | 15 | 3.5 | 49 | 4.6 |
| White Sucker | Catostomus commersoni | 7 | 14.3 | 21 | 11.2 | 47 | 11.7 | 89 | 20.7 | 164 | 15.4 |
| Yellow Perch | Perca flavescens | 0 | 0.0 | 3 | 1.6 | 0 | 0.0 | 0 | 0.0 | 3 | 0.3 |
| Total |  | 49 | 100 | 187 | 100 | 401 | 100 | 430 | 100 | 1067 | 100 |

Table 4: Lake Sturgeon catch-per-unit effort (CPUE; \# fish/100 m net/24 h) by location and zone, for gill nets set during juvenile Lake Sturgeon monitoring, fall, 2018.

| Location | Zone | \# of Sites | Effort (gillnet hours) | \# of Lake Sturgeon | $\begin{gathered} \text { CPUE } \\ \text { (\#LKST/100m/24h) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Burntwood River | BWR-A | 6 | 139.9 | 6 | 1.03 |
|  | BWR-B | 5 | 108.0 | 2 | 0.44 |
|  | BWR-C | 8 | 178.4 | 3 | 0.40 |
| Total |  | 19 | 426.3 | 11 | 0.62 |
| Split Lake | SPL-A | 21 | 606.9 | 57 | 2.25 |
| Total |  | 21 | 606.9 | 57 | 2.25 |
| Future Keeyask Reservoir | BR-D | 4 | 135.2 | 5 | 0.89 |
|  | GL-A | 11 | 314.8 | 17 | 1.30 |
|  | GL-B | 18 | 480.1 | 60 | 3.00 |
|  | GL-C | 17 | 446.7 | 68 | 3.65 |
| Total |  | 50 | 1376.8 | 150 | 2.61 |
| Stephens Lake | STL-A | 17 | 576.5 | 19 | 0.79 |
|  | STL-B | 32 | 1022.5 | 55 | 1.29 |
| Total |  | 49 | 1599.0 | 74 | 1.11 |

 rows indicate construction monitoring.

| Location | Year | Start Date | Completion Date | Mesh Size | \# Sites | Effort (gillnet hrs ${ }^{\text {a }}$ ) | \# Lake Sturgeon ${ }^{\text {b }}$ | CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Upper Split Lake Area |  |  |  |  |  |  |  |  |
| Burntwood River | 2012 | 29-Aug | 08-Sep | 1" - 6" | 37 | 767 | 33 | 1.03 |
|  | 2014 | 08-Sep | 16-Sep | 1"-6" | 28 | 734 | 42 | 1.37 |
|  | 2015 | 29-Aug | 04-Oct | 1"-6" | 28 | 858 | 35 | 0.78 |
|  | 2016 | 07-Sep | 18-Sep | 1"-6" | 24 | 594 | 26 | 1.05 |
|  | 2017 | 06-Sep | 12-Sep | 1"-6" | 24 | 660 | 34 | 1.24 |
|  | 2018 | 09-Sep | 20-Sep | 1"-6" | 19 | 426 | 11 | 0.62 |
| Kelsey GS Area ${ }^{\text {c }}$ | 2015 | 29-Aug | 04-Oct | 1"-6" | 7 | 248 | 7 | 0.68 |
|  | 2016 | 07-Sep | 18-Sep | 1" -6" | 9 | 203 | 8 | 0.95 |
|  | 2017 | 14-Sep | 15-Sep | 1"-6" | 10 | 232 | 6 | 0.62 |
| Split Lake | 2015 | 29-Aug | 04-Oct | 1"-6" | 9 | 192 | 9 | 1.13 |
|  | 2016 | 07-Sep | 18-Sep | 1"-6" | 7 | 193 | 6 | 0.75 |
|  | 2017 | 05-Sep | 13-Sep | 1"-6" | 8 | 175 | 19 | 2.60 |
|  | 2018 | 09-Sep | 20-Sep | 1"-6" | 21 | 607 | 57 | 2.25 |
| Future Keeyask Reservoir ${ }^{\text {d }}$ |  |  |  |  |  |  |  |  |
|  | 2007 | 28-Sep | 03-Oct | 8mm - 5" | 26 | 165 | 0 | 0 |
|  | 2008 | 12-Sep | 27-Sep | 1.5 "- 8" | 15 | 3072 | 126 | 0.98 |
|  | 2010 | 21-Sep | 29-Sep | 1"-5" | 27 | 851 | 69 | 1.95 |
|  | 2011 | 18-Sep | 24-Sep | 1"-5" | 25 | 662 | 121 | 4.39 |
|  | 2012 | 29-Aug | 09-Sep | 1"-6" | 30 | 745 | 101 | 3.25 |
|  | 2014 | 08-Sep | 16-Sep | 1"-6" | 30 | 765 | 112 | 3.51 |
|  | 2015 | 11-Sep | 20-Sep | 1"-6" | 34 | 912 | 139 | 3.66 |
|  | 2016 | 12-Sep | 23-Sep | 1"-6" | 37 | 997 | 96 | 2.31 |
|  | 2017 | 09-Sep | 19-Sep | 1"-6" | 51 | 1551 | 177 | 2.74 |
|  | 2018 | 09-Sep | 19-Sep | 1" - 6" | 50 | 1377 | 150 | 2.61 |
|  |  |  |  |  |  |  |  |  |
| Stephens Lake | 2007 | 19-Sep | 23-Sep | 2"-5" | 15 | 48 | 0 | 0 |
|  | 2008 | 11-Sep | 18-Sep | 3.75"-8" | 12 | 295 | 8 | 0.65 |
|  | 2009 | 14-Sep | 20-Sep | $1.5{ }^{\prime \prime}-5^{\prime \prime}$ | 18 | 634 | 23 | 0.87 |
|  | 2010 | 22-Sep | 29-Sep | 1"-5" | 18 | 611 | 32 | 1.26 |
|  | 2011 | 21-Sep | 01-Oct | 1"-5" | 30 | 974 | 37 | 0.91 |
|  | 2012 | 11-Sep | 23-Sep | 1"-6" | 19 | 1193 | 87 | 1.75 |
|  | 2014 | 18-Sep | 28-Sep | 1"-6" | 94 | 921 | 47 | 1.23 |
|  | 2015 | 22-Sep | 02-Oct | 1"-6" | 44 | 1154 | 54 | 1.12 |
|  | 2016 | 12-Sep | 23-Sep | 1"-6" | 37 | 1384 | 66 | 1.14 |
|  | 2017 | 09-Sep | 19-Sep | 1"-6" | 40 | 1796 | 148 | 1.98 |
|  | 2018 | 09-Sep | 21-Sep | 1"-6" | 49 | 1599 | 74 | 1.11 |

a - Gillnet set durations were standardized to 100 m of net and then summed to calculate the total gillnet hours for each study.
b - Does not include Lake Sturgeon recaptured more than once in the same study.
c - Removed from sampling in 2018
d - Birthday Rapids to Gull Rapids.
 indicate cohorts not present in the corresponding study year. The Kelsey GS area was not samples and more sampling sites were added to Split Lake in 2018 in an attempt to locate hatchery-reared fish stocked in the Burntwood River.

| Location | Cohort Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| Upper Split Lake Area |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Burntwood River |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 Study Year | 0 | 2 | 5 | 3 | 5 | 15 | 6 | 0 | 2 | 1 | 4 | 0 | 0 | 0 | - | - | - | - | - | - | - |
| 2012 Study Year | 0 | 2 | 1 | 4 | 0 | 4 | 0 | 1 | 5 | 3 | 1 | 0 | 3 | 7 | 1 | - | - | - | - | - | - |
| 2015 Study Year | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 1 | 3 | 1 | 2 | 0 | 0 | 5 | 4 | 4 | 0 | 0 | - | - | - |
| 2016 Study Year | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 4 | 5 | 0 | 7 | 2 | 0 | 1 | - | - |
| 2017 Study Year | 0 | 1 | 0 | 0 | 1 | 2 | 1 | 0 | 0 | 0 | 2 | 1 | 5 | 2 | 0 | 2 | 3 | 1 | 7 | 3 | - |
| 2018 Study Year | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 2 | 2 | 1 | 0 |
| Total | 0 | 7 | 6 | 8 | 7 | 23 | 9 | 2 | 10 | 5 | 10 | 1 | 14 | 19 | 5 | 14 | 5 | 3 | 10 | 4 | 0 |
| Present in the Catch | No | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | No |
| Split Lake |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 4 | 1 | 0 | - | - | - |
| 2016 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | - | - |
| 2017 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 3 | 1 | 3 | 1 | 0 | 2 | 0 | 0 | 3 | 0 | - |
| 2018 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 2 | 1 | 1 | 0 | 3 | 9 | 1 | 26 | 2 | 1 | 2 | 1 | 0 |
| Total | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 3 | 3 | 6 | 1 | 7 | 11 | 1 | 32 | 4 | 2 | 5 | 1 | 0 |
| Present in the Catch | No | No | No | No | No | Yes | No | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | No |
| Kelsey GS |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | - | - | - |
| 2016 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | - | - |
| 2017 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | - |
| Total | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 1 | 0 | 2 | 0 | 2 | 5 | 1 | 1 | 0 | 0 | 0 | 0 | - |
| Present in the Catch | No | No | No | No | No | Yes | Yes | Yes | Yes | No | Yes | No | Yes | Yes | Yes | Yes | No | No | No | No | - |
| Future Keeyask Reservoir |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2008 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 2 | 14 | - | - | - | - | - | - | - | - | - | - |
| 2010 Study Year | 1 | 0 | 1 | 0 | 0 | 6 | 3 | 1 | 3 | 5 | 18 | 0 | 0 | - | - | - | - | - | - | - | - |
| 2011 Study Year | 0 | 1 | 0 | 0 | 0 | 5 | 2 | 2 | 7 | 5 | 94 | 1 | 2 | 0 | - | - | - | - | - | - | - |
| 2012 Study Year | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 2 | 12 | 6 | 60 | 3 | 1 | 4 | 0 | - | - | - | - | - | - |
| 2014 Study Year | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 6 | 2 | 58 | 3 | 4 | 7 | 3 | 9 | 0 | - | - | - | - |
| 2015 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 10 | 7 | 71 | 1 | 1 | 3 | 6 | 11 | 3 | 4 | - | - | - |
| 2016 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 15 | 0 | 29 | 2 | 1 | 5 | 6 | 13 | 6 | 4 | 4 | - | - |
| 2017 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 6 | 3 | 56 | 2 | 2 | 11 | 7 | 20 | 10 | 10 | 10 | 1 | - |
| 2018 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 33 | 5 | 3 | 6 | 4 | 9 | 5 | 9 | 34 | 5 | 1 |
| Total | 1 | 1 | 1 | 1 | 0 | 15 | 9 | 10 | 74 | 34 | 433 | 17 | 14 | 36 | 26 | 62 | 24 | 27 | 48 | 6 | 1 |
| Present in the Catch | Yes | Yes | Yes | Yes | No | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes |

# Number of wild Lake Sturgeon captured from 2008 to 2018, from which ages and cohorts were determined. Grey highlighted rows indicate construction monitoring and red values indicate cohorts not present in the 

 corresponding study year (continued).| Location | Cohort Year |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| Stephens Lake |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | - |
| 2009 Study Year | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 3 | 1 | 0 | 2 | 0 | - | - | - | - | - | - | - |  | - |
| 2010 Study Year | 0 | 0 | 0 | 0 | 1 | 3 | 0 | 1 | 5 | 7 | 14 | 0 | 0 | - | - | - | - | - | - |  | - |
| 2011 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 28 | 2 | 0 | 1 | - | - | - | - | - |  | - |
| 2012 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 4 | 49 | 1 | 2 | 2 | 0 | - | - | - | - |  | - |
| 2014 Study Year | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 5 | 4 | 25 | 1 | 4 | 5 | 0 | 0 | 0 | - | - |  | - |
| 2015 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 3 | 19 | 1 | 1 | 3 | 0 | 4 | 2 | 11 | - |  | - |
| 2016 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 4 | 31 | 0 | 0 | 2 | 1 | 3 | 4 | 8 | 0 |  | - |
| 2017 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 19 | 2 | 0 | 3 | 0 | 11 | 4 | 20 | 9 | 5 | - |
| 2018 Study Year | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 4 | 1 | 9 | 3 | 20 | 4 | 3 | 0 |
| Total | 0 | 0 | 1 | 1 | 1 | 5 | 3 | 4 | 26 | 24 | 191 | 7 | 7 | 20 | 2 | 27 | 13 | 59 | 13 | 8 | 0 |
| Present in the Catch | No | No | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | Yes | No |

Table 7: Mean length, weight, and condition factor of Lake Sturgeon captured during juvenile Lake Sturgeon monitoring, fall 2018.

| Waterbody | Fork Length (mm) |  |  |  | Weight (g) |  |  |  | Condition Factor |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{n}^{\text {a }}$ | Mean | Std ${ }^{\text {b }}$ | Range | n | Mean | Std | Range | n | Mean | Std | Range |
| Upper Split Lake Area |  |  |  |  |  |  |  |  |  |  |  |  |
| Burntwood River | 11 | 455 | 191 | 205-764 | 11 | 950 | 1,043 | 25-3,000 | 11 | 0.62 | 0.23 | 0.20-0.87 |
| Split Lake | 57 | 583 | 151 | 230-996 | 57 | 1,816 | 1,428 | 25-7,350 | 57 | 0.75 | 0.14 | 0.21-1.46 |
|  | 68 | 562 | 164 | 205-996 | 68 | 1,676 | 1,403 | 25-7,350 | 68 | 0.73 | 0.17 | 0.20-1.46 |
| Future Keeyask Reservoir |  |  |  |  |  |  |  |  |  |  |  |  |
| Wild | 128 | 524 | 206 | 87-1,031 | 127 | 1,565 | 1,634 | 50-8,500 | 127 | 0.73 | 0.13 | 0.32-1.30 |
| Hatchery | 17 | 396 | 57 | 255-479 | 17 | 394 | 148 | 100-700 | 17 | 0.60 | 0.05 | 0.53-0.72 |
|  | $150{ }^{\text {c }}$ | 504 | 197 | 87-1,031 | $149{ }^{\text {c }}$ | 1,391 | 1,566 | 50-8,500 | $149{ }^{\text {c }}$ | 0.71 | 0.13 | 0.32-1.30 |
| Stephens Lake |  |  |  |  |  |  |  |  |  |  |  |  |
| Wild | 52 | 488 | 158 | 222-837 | 52 | 1,172 | 1,254 | 50-4,925 | 52 | 0.72 | 0.10 | 0.46-0.90 |
| Hatchery | 17 | 432 | 64 | 346-503 | 17 | 596 | 239 | 275-900 | 17 | 0.69 | 0.08 | 0.48-0.85 |
|  | $74{ }^{\text {d }}$ | 469 | 139 | 222-837 | $74{ }^{\text {d }}$ | 994 | 1,092 | 50-4,925 | $74{ }^{\text {d }}$ | 0.71 | 0.09 | 0.46-0.90 |

a - Number of fish measured.
b - Standard deviation.
c - Total includes an additional 5 fish that were not conclusively identified as either wild or hatchery fish.
d - Total includes an additional 5 fish that were not conclusively identified as either wild or hatchery fish.

Table 8: Recapture summary for wild Lake Sturgeon caught in the Keeyask Study Area between 2008 and 2018.

| Recapture Location | Sampling Year | Tagging Location |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Upper Split Lake Area | Future Keeyask Reservoir | Stephens Lake |
|  |  | $\mathrm{n}^{\text {a }}$ | $\mathrm{n}^{\text {a }}$ | $\mathrm{n}^{\text {a }}$ |
| Upper Split Lake Area | 2011 | 0 | 0 | 0 |
|  | 2012 | 2 | 0 | 0 |
|  | 2014 | 2 | 0 | 0 |
|  | 2015 | 2 | 0 | 0 |
|  | 2016 | 2 | 0 | 0 |
|  | 2017 | 3 | 0 | 0 |
|  | 2018 | 4 | 0 | 0 |
| Future Keeyask Reservoir | 2008 | 0 | 9 | 0 |
|  | 2010 | 0 | 2 | 0 |
|  | 2011 | 0 | 4 | 0 |
|  | 2012 | 0 | 8 | 0 |
|  | 2014 | 0 | 17 | 0 |
|  | 2015 | 0 | 20 | 0 |
|  | 2016 | 0 | 11 | 0 |
|  | 2017 | 0 | 17 | 0 |
|  | 2018 | 0 | 18 | 0 |
| Stephens Lake | 2009 | 0 | 0 | 0 |
|  | 2010 | 0 | 0 | 0 |
|  | 2011 | 0 | 0 | 0 |
|  | 2012 | 0 | 0 | 11 |
|  | 2014 | 0 | 0 | 8 |
|  | 2015 | 0 | 0 | 7 |
|  | 2016 | 0 | 0 | 14 |
|  | 2017 | 0 | 3 | 17 |
|  | 2018 | 0 | 1 | 10 |

a - Number of Lake Sturgeon.

Table 9: $\quad$ Number ( $n$ ) and percentage (\%) of catch of hatchery-reared Lake Sturgeon caught in the Keeyask Study Area between 2014 and 2018.

| Capture Location | Sample Year | Release Location |  |  |  |  |  | Total | \% of Total Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Burntwood River |  | Future Keeyask Reservoir |  | Stephens Lake |  |  |  |
|  |  | n | \% of Catch | n | \% of Catch | n | \% of Catch |  |  |
| Upper Split Lake Area | 2014 | 1 | 2.4 | 0 | 0.0 | 0 | 0.0 | 1 | 2.4 |
|  | 2015 | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
|  | 2016 | 1 | 2.5 | 0 | 0.0 | 0 | 0.0 | 1 | 2.5 |
|  | 2017 | 3 | 5.1 | 0 | 0.0 | 0 | 0.0 | 3 | 5.1 |
|  | 2018 | 1 | 1.8 | 0 | 0.0 | 0 | 0.0 | 1 | 1.8 |
| Future Keeyask Reservoir | 2014 | 1 | 0.9 | 0 | 0.0 | 0 | 0.0 | 1 | 0.9 |
|  | 2015 | 1 | 0.7 | 2 | 1.4 | 0 | 0.0 | 3 | 2.2 |
|  | 2016 | 0 | 0.0 | 7 | 7.3 | 0 | 0.0 | 7 | 7.3 |
|  | 2017 | 1 | 0.6 | 20 | 11.6 | 0 | 0.0 | 21 | 11.9 |
|  | 2018 | 1 | 0.7 | 16 | 10.7 | 0 | 0.0 | 17 | 11.3 |
| Stephens Lake | 2014 | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 |
|  | 2015 | 0 | 0.0 | 0 | 0.0 | 4 | 7.4 | 4 | 8.5 |
|  | 2016 | 0 | 0.0 | 1 | 1.5 | 4 | 6.1 | 5 | 7.6 |
|  | 2017 | 0 | 0.0 | 11 | 7.4 | 40 | 27.0 | 51 | 34.5 |
|  | 2018 | 0 | 0.0 | 3 | 4.1 | 14 | 18.9 | 17 | 23.0 |

Table 10: Catch-per-unit-effort (CPUE; \# fish/100 m net/24 h) for hatchery and wild caught Lake Sturgeon in Stephens Lake and the future Keeyask reservoir, Fall 2018.

| Location | Effort (gillnet hours) | \# of Lake Sturgeon | CPUE <br> (\#LKST/100m/24h) |
| :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir |  |  |  |
| Wild | $1,376.8$ | 128 | 2.23 |
| Hatchery | $1,376.8$ | 17 | 0.30 |
| Stephens Lake | Total | $\mathbf{1 5 0}$ | $\mathbf{2 . 6 1}$ |
| Wild | $1,599.0$ |  | 0.78 |
| Hatchery | $1,599.0$ | 52 | 0.26 |
|  | Total | 17 | $\mathbf{1 . 1 1}$ |

a - total includes an additional 5 fish that were not conclusively identified as either wild or hatchery fish.
b - total includes an additional 5 fish that were not conclusively identified as either wild or hatchery fish.

Table 11: Number and ages of hatchery-reared Lake Sturgeon released as age-1 fish and captured during juvenile Lake Sturgeon studies since 2014.

| Monitoring Year | Capture Location |  |  |
| :---: | :---: | :---: | :---: |
|  | Burntwood River | Future Keeyask Reservoir | Stephens Lake |
| 2014 | $\begin{gathered} 1 \\ \text { (1 year old) } \end{gathered}$ | $\begin{gathered} 1 \\ \text { (1 year old) } \end{gathered}$ | - |
| 2015 | - | (2 were 1 year old) <br> (1 was 2 years old) | (All were 1 year old) |
| 2016 | $\begin{gathered} 1 \\ (3 \text { years old) } \end{gathered}$ | 7 (All were 2 years old) | (All were 2 years old) |
| 2017 | 3 (All were 4 years old) | 21 <br> (9 were 1 year old) <br> (11 were 3 years old) <br> (1 was 4 years old) | 51 (33 were 1 year old) (18 were 3 years old) |
| 2018 | (1 was 5 years old) | $\begin{gathered} 17 \\ (1 \text { was } 1 \text { years old) } \\ \text { ( } 8 \text { were } 2 \text { years old) } \\ \text { ( } 8 \text { were } 4 \text { years old) } \end{gathered}$ | 17 (7 were 2 years old) (10 were 4 years old) |

a - fish released in the Burntwood River but caught in Split Lake.

## FIGURES



Figure 1:
Cohort frequency distribution by zone, for all aged Lake Sturgeon captured in the Burntwood River, fall 2018. Cohorts prior to 2009 (i.e., age-9 fish) are not fully represented as ageing structures are not collected from fish > $\mathbf{8 0 0} \mathbf{~ m m}$ fork length (indicated by vertical dashed line).


Fork Length Interval (mm)

Figure 2: Length-frequency distributions for Lake Sturgeon captured in gill nets set in the Upper Split Lake Area: A) the Burntwood River and B) Split Lake, fall 2018.


Figure 3: Comparison of weight (g) at-fork length (mm) (log transformed) for Lake Sturgeon captured in: A) the Upper Split Lake Area B) the future Keeyask reservoir and C) Stephens Lake, fall 2018. Only one hatchery-reared Lake Sturgeon was captured in Split Lake therefore hatchery and wild fish were not analysed separately in the Upper Split Lake Area (A).


Figure 4: Mean condition factor-at-age for Lake Sturgeon captured in the Upper Split Lake Area during baseline studies (2011-2013) and during the construction period (2014-2018). Error bars represent standard deviations. There were not enough baseline data to allow for significance testing. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish $\mathbf{> 8 0 0} \mathbf{~ m m}$ fork length, which corresponds to fish older than age-9).


Figure 5: Mean fork length-at-age for Lake Sturgeon captured in the Upper Split Lake Area during baseline studies (20112013) and the construction period (2014-2018). Error bars represent standard deviations. There were not enough baseline or construction data to allow for significance testing. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish $\mathbf{> 8 0 0} \mathbf{~ m m}$ fork length, which corresponds to fish older than age-9).


Cohort Year

Figure 6: Cohort frequency distributions for all aged Lake Sturgeon captured in Zone SPL-A of Split Lake, fall 2018. A single hatchery-reared fish captured in Zone SPL-A was included as part of the 2013 cohort. Cohorts prior to 2009 (i.e., age-9 fish) are not fully represented as ageing structures are not collected from fish $\mathbf{> 8 0 0} \mathbf{m m}$ fork length (indicated by vertical dashed line).


Figure 7: Cohort frequency distributions for all aged Lake Sturgeon captured by zone in the future Keeyask reservoir (A) and by hatchery and wild Lake Sturgeon (B), fall 2018. Cohorts prior to 2009 (i,e., age-9 fish) are not fully represented as ageing structures are not collected from fish $>\mathbf{8 0 0} \mathbf{~ m m}$ fork length (indicated by vertical dashed line).


Figure 8. Juvenile Lake Sturgeon abundance (i.e., fish < $\mathbf{8 0 0} \mathbf{~ m m}$ fork length) estimates based on POPAN best model for the future Keeyask reservoir (2010-2018). Horizontal line inside the box represents the estimated abundance (i.e., the number of juvenile Lake Sturgeon), the red dots represent the min and max estimates, and the vertical bar lines represent the upper and lower $95 \%$ confidence intervals.


Figure 9: Mean condition factor-at-age for wild Lake Sturgeon captured in the future Keeyask reservoir during baseline studies (2008-2013) and the construction period (2014-2018). Letters denote significant differences between groups ( $\mathbf{t}$-test and Mann Whitney U test, $\boldsymbol{p}<\mathbf{0 . 0 5}$ ). Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish >800 mm fork length, which corresponds to fish older than age-9).


Figure 10: Mean fork length-at-age for wild Lake Sturgeon captured in the future Keeyask reservoir during baseline studies (2008-2013) and the construction period (2014-2018). Letters denote significant differences between groups (ttest and Mann Whitney $U$ test, $\boldsymbol{p} \boldsymbol{< 0 . 0 5 )}$. Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish $\mathbf{> 8 0 0} \mathbf{~ m m}$ fork length, which corresponds to fish older than age-9).

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Fork Length Interval (mm)

Figure 11: Fork length frequency distributions for Lake Sturgeon captured in gill nets set in: A) the future Keeyask reservoir and B) Stephens Lake, fall 2018.


Figure 12: Fork length-at-age (A), weight-at-age (B), and condition factor-at-age (B) for wild (blue bars) and hatchery (red bars) Lake Sturgeon caught in the future Keeyask reservoir since 2014. Letters denote significant differences between groups (Mann Whitney U test, p < 0.05). Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations.


Figure 13. Fork length-at-age (A) and von Bertalanffy growth curve analysis (B) for all Lake Sturgeon caught during baseline (red; 2008-2012) and construction (blue; 2014-2018) monitoring years in the future Keeyask reservoir. Fish older that age-9 were not included in the analysis as they are not fully represented in the catch (ageing structures are not collected from fish >800 mm fork length, which corresponds to fish older than age-9).


Figure 14. Fork length-at-age (A) and von Bertalanffy growth curve analysis (B) for all wild (blue) and hatchery-reared (red) Lake Sturgeon released and/or recaptured in the future Keeyask reservoir since stocking began in 2014. Fish older that age-9 were not included in the analysis as they are not fully represented in the catch (ageing structures are not collected from fish >800 mm fork length, which corresponds to fish older than age-9).


Figure 15: Cohort frequency distributions for all aged Lake Sturgeon captured in Stephens Lake by zone (A) and by hatchery and wild Lake Sturgeon (B), fall 2018. Cohorts prior to 2009 (i.e., age-9 fish) are not fully represented as ageing structures are not collected from fish $\mathbf{> 8 0 0} \mathbf{~ m m}$ fork length (indicated by vertical dashed line).

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Figure 16. Juvenile Lake Sturgeon abundance estimates based on POPAN best model for Stephens Lake (2010-2018). Results of the POPAN abundance estimate are presented in black. Horizontal line inside the box represents the estimated abundance (i.e., the number of juvenile Lake Sturgeon), the red dots represent the min and max estimates, and the vertical bar lines represent the upper and lower $95 \%$ confidence intervals.


Figure 17: Mean condition factor-at-age for Lake Sturgeon captured in Stephens Lake during baseline studies (2008-2013) and the construction period (2014-2018). Letters denote significant differences between groups (t-test and Mann Whitney U test, $\boldsymbol{p}<\mathbf{0 . 0 5}$ ). Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish $\mathbf{>} \mathbf{8 0 0} \mathbf{~ m m}$ fork length, which corresponds to fish older than age-9).

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Figure 18: Mean fork length-at-age for wild Lake Sturgeon captured in Stephens Lake during baseline studies (2008-2013) and the construction period (2014-2018). Letters denote significant differences between groups (t-test and Mann Whitney U test, $\boldsymbol{p}<\mathbf{0 . 0 5}$ ). Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations. Fish older that age-9 were not included as they are not fully represented in the catch (ageing structures are not collected from fish $\mathbf{>} \mathbf{8 0 0} \mathbf{~ m m}$ fork length, which corresponds to fish older than age-9).

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Figure 19: Fork length-at-age (A), weight-at-age (B), and condition factor-at-age (B) for wild (blue bars) and hatchery (red bars) Lake Sturgeon caught in Stephens Lake since 2014. Letters denote significant differences between groups (t-test and Mann Whitney U test, p < 0.05). Statistical comparisons only conducted where sample sizes were greater than eight individuals. Error bars represent standard deviations.


Figure 20. Fork length-at-age (A) and von Bertalanffy growth curve analysis (B) for all wild (blue) and hatchery-reared (red) Lake Sturgeon released and/or recaptured in Stephens Lake since stocking began in 2014. Fish older that age-9 were not included in the analysis as they are not fully represented in the catch (ageing structures are not collected from fish >800 mm fork length, which corresponds to fish older than age-9).

## MAPS



Map 1: $\quad$ Map of Nelson River showing the site of Keeyask Generating Station and the juvenile Lake Sturgeon population monitoring study setting.


Satellite Imagery - October 12th, 201
Map 2:
Map of instream structures at the Keeyask Generating Station site, October, 2018.


Map 3:
Map of sites fished with gill nets in the Upper Split Lake Area (Burntwood River and Split Lake), fall 2018.

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Map 4:
Map of Lake Sturgeon yearling stocking sites in the Burntwood River since 2014.

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Map 5:
Map of sites fished with gill nets in the future Keeyask reservoir, fall 2018.


Map 6:
Map of Lake Sturgeon yearling stocking sites in the future Keeyask reservoir since 2014.

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Map 7:
Map of sites fished with gill nets in Stephens Lake, fall 2018.

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Map 8:
Map of Lake Sturgeon yearling stocking sites in Stephens Lake since 2014.

## APPENDICES

# APPENDIX 1: <br> LOCATIONS AND SITE-SPECIFIC PHYSICAL MEASUREMENTS COLLECTED AT GILLNETTING SITES, FALL 2018. 

Table A1-1: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the Upper Split Lake Area, fall 201877
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Table A1-3: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in Stephens Lake, fall 2018. ..... 82

Table A1-1: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the Upper Split Lake Area, fall 2018. Sites set in each region are indicated as follows Burntwood River (BWR) and Split Lake (SPL).

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-01 | SPL-A | 658162 | 6221849 | 9-Sep-18 | 16.0 | 11-Sep-18 | 14.5 | 41.77 | 12.3 | 14.1 |
| GN-02 | SPL-A | 658527 | 6221941 | 9-Sep-18 | 16.0 | 11-Sep-18 | 14.5 | 42.68 | 11.2 | 10.9 |
| GN-03 | SPL-A | 659125 | 6222047 | 9-Sep-18 | 16.0 | 11-Sep-18 | 14.5 | 43.33 | 13.3 | 11.0 |
| GN-04 | SPL-A | 660038 | 6222051 | 9-Sep-18 | 16.0 | 11-Sep-18 | 14.5 | 43.95 | 9.5 | 7.1 |
| GN-05 | SPL-A | 660458 | 6221950 | 11-Sep-18 | 14.5 | 12-Sep-18 | 14.5 | 21.35 | 8.9 | 5.5 |
| GN-06 | SPL-A | 660913 | 6221977 | 11-Sep-18 | 14.5 | 12-Sep-18 | 14.5 | 20.48 | 8.7 | 6.2 |
| GN-07 | SPL-A | 661481 | 6221993 | 11-Sep-18 | 14.5 | 12-Sep-18 | 14.5 | 19.68 | 9.2 | 5.7 |
| GN-08 | SPL-A | 661772 | 6221876 | 11-Sep-18 | 14.5 | 12-Sep-18 | 14.5 | 19.07 | 11.8 | 10.0 |
| GN-09 | BWR-A | 630582 | 6214209 | 12-Sep-18 | 14.5 | 13-Sep-18 | 14.5 | 20.28 | 5.0 | 9.8 |
| GN-10 | BWR-A | 630849 | 6214517 | 12-Sep-18 | 14.5 | 13-Sep-18 | 14.5 | 20.87 | 5.8 | 10.2 |
| GN-11 | BWR-A | 631203 | 6214878 | 12-Sep-18 | 14.5 | 13-Sep-18 | 14.5 | 21.45 | 11.3 | 12.2 |
| GN-12 | BWR-A | 631509 | 6215121 | 12-Sep-18 | 14.5 | 13-Sep-18 | 14.5 | 22.03 | 10.0 | 6.4 |
| GN-13 | BWR-A | 632711 | 6215939 | 13-Sep-18 | 14.5 | 14-Sep-18 | 14.5 | 18.72 | 12.0 | 13.9 |
| GN-14 | BWR-A | 633813 | 6216352 | 13-Sep-18 | 14.5 | 14-Sep-18 | 14.5 | 19.05 | 11.7 | 11.9 |
| GN-15 | BWR-B | 635186 | 6219049 | 13-Sep-18 | 14.5 | 14-Sep-18 | 14.5 | 19.52 | 10.7 | 10.0 |
| GN-16 | BWR-B | 633107 | 6218234 | 13-Sep-18 | 14.5 | 14-Sep-18 | 14.5 | 18.88 | 9.1 | 8.3 |
| GN-17 | BWR-B | 631753 | 6218195 | 14-Sep-18 | 14.5 | 15-Sep-18 | 14.5 | 18.47 | 10.1 | 11.2 |
| GN-18 | BWR-B | 630613 | 6219142 | 14-Sep-18 | 14.5 | 15-Sep-18 | 14.5 | 18.73 | 10.6 | 11.2 |
| GN-19 | BWR-B | 631770 | 6220142 | 14-Sep-18 | 14.5 | 15-Sep-18 | 14.5 | 18.88 | 8.6 | 11.2 |
| GN-20 | BWR-C | 632714 | 6220975 | 14-Sep-18 | 14.5 | 15-Sep-18 | 14.5 | 19.33 | 10.1 | 10.5 |
| GN-21 | BWR-C | 636447 | 6223001 | 15-Sep-18 | 14.5 | 16-Sep-18 | 13.5 | 19.50 | 10.0 | 12.7 |
| GN-22 | BWR-C | 638038 | 6223836 | 15-Sep-18 | 14.5 | 16-Sep-18 | 13.5 | 20.72 | 10.1 | 10.5 |

Table A1-1: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the Upper Split Lake Area, fall 2018. Sites set in each region are indicated as follows Burntwood River (BWR) and Split Lake (SPL) (continued).

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-23 | BWR-C | 638899 | 6224266 | 15-Sep-18 | 14.5 | 16-Sep-18 | 13.5 | 21.25 | 13.5 | 12.7 |
| GN-24 | BWR-C | 639896 | 6224640 | 15-Sep-18 | 14.5 | 16-Sep-18 | 13.5 | 21.63 | 8.8 | 12.4 |
| GN-25 | BWR-C | 648476 | 6224358 | 16-Sep-18 | 13.5 | 17-Sep-18 | 13.0 | 17.73 | 5.4 | 5.5 |
| GN-26 | BWR-C | 648522 | 6224225 | 16-Sep-18 | 13.5 | 17-Sep-18 | 13.0 | 17.87 | 6.7 | 7.3 |
| GN-27 | BWR-C | 650393 | 6224088 | 16-Sep-18 | 13.5 | 17-Sep-18 | 13.0 | 18.02 | 3.6 | 6.0 |
| GN-28 | SPL-A | 651682 | 6223960 | 16-Sep-18 | 13.5 | 17-Sep-18 | 13.0 | 18.32 | 5.9 | 6.2 |
| GN-29 | SPL-A | 663081 | 6225306 | 17-Sep-18 | 13.0 | 18-Sep-18 | 13.0 | 19.33 | 10.9 | 12.1 |
| GN-30 | SPL-A | 663420 | 6224966 | 17-Sep-18 | 13.0 | 18-Sep-18 | 13.0 | 19.50 | 10.3 | 10.4 |
| GN-31 | SPL-A | 664072 | 6225094 | 17-Sep-18 | 13.0 | 18-Sep-18 | 13.0 | 19.95 | 12.2 | 10.4 |
| GN-32 | SPL-A | 664545 | 6225463 | 17-Sep-18 | 13.0 | 18-Sep-18 | 13.0 | 20.28 | 12.3 | 10.2 |
| GN-33 | SPL-A | 665046 | 6225865 | 18-Sep-18 | 13.0 | 19-Sep-18 | 12.0 | 23.18 | 9.4 | 9.7 |
| GN-34 | SPL-A | 665349 | 6225459 | 18-Sep-18 | 13.0 | 19-Sep-18 | 12.0 | 22.63 | 9.5 | 8.8 |
| GN-35 | SPL-A | 665544 | 6225245 | 18-Sep-18 | 13.0 | 19-Sep-18 | 12.0 | 22.12 | 9.0 | 10.0 |
| GN-36 | SPL-A | 665979 | 6225051 | 18-Sep-18 | 13.0 | 19-Sep-18 | 12.0 | 21.75 | 10.2 | 9.2 |
| GN-37 | SPL-A | 659296 | 6223392 | 19-Sep-18 | 12.0 | 20-Sep-18 | 11.5 | 22.30 | 5.3 | 4.6 |
| GN-38 | SPL-A | 658973 | 6223049 | 19-Sep-18 | 12.0 | 20-Sep-18 | 11.5 | 22.78 | 5.3 | 5.8 |
| GN-39 | SPL-A | 657458 | 6222149 | 19-Sep-18 | 12.0 | 20-Sep-18 | 11.5 | 23.08 | 10.3 | 11.1 |
| GN-40 | SPL-A | 656797 | 6222205 | 19-Sep-18 | 12.0 | 20-Sep-18 | 11.5 | 23.48 | 4.8 | 7.9 |

Table A1-2: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the future Keeyask reservoir, fall 2018.

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-01 | GL-C | 356639 | 6248179 | 9-Sep-18 | 12.0 | 10-Sep-18 | 11.0 | 18.33 | 10.9 | 10.1 |
| GN-02 | GL-C | 356053 | 6247789 | 9-Sep-18 | 12.0 | 10-Sep-18 | 11.0 | 18.72 | 6.7 | 7.3 |
| GN-03 | GL-B | 353958 | 6244298 | 9-Sep-18 | 12.0 | 10-Sep-18 | 11.0 | 19.08 | 10.4 | 12.5 |
| GN-04 | GL-C | 356500 | 6248051 | 10-Sep-18 | 11.0 | 11-Sep-18 | 11.0 | 25.22 | 6.8 | 6.3 |
| GN-05 | GL-C | 355513 | 6247296 | 10-Sep-18 | 11.0 | 11-Sep-18 | 11.0 | 24.17 | 7.0 | 6.9 |
| GN-06 | GL-B | 354436 | 6244711 | 10-Sep-18 | 11.0 | 11-Sep-18 | 11.0 | 22.72 | 13.4 | 11.4 |
| GN-07 | GL-B | 353664 | 6244170 | 11-Sep-18 | 11.0 | 12-Sep-18 | 11.0 | 28.17 | 11.2 | 11.7 |
| GN-08 | GL-B | 354346 | 6244502 | 11-Sep-18 | 11.0 | 12-Sep-18 | 11.0 | 27.83 | 10.7 | 12.2 |
| GN-09 | GL-C | 355336 | 6247142 | 11-Sep-18 | 11.0 | 12-Sep-18 | 11.0 | 28.25 | 11.5 | 8.6 |
| GN-10 | GL-C | 356439 | 6248037 | 11-Sep-18 | 11.0 | 12-Sep-18 | 11.0 | 28.42 | 7.1 | 7.9 |
| GN-11 | GL-A | 349351 | 6244313 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 20.58 | 10.3 | 11.2 |
| GN-12 | GL-A | 348408 | 6244278 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 19.25 | 12.6 | 12.1 |
| GN-13 | GL-B | 353678 | 6244041 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 20.75 | 8.0 | 9.9 |
| GN-14 | GL-B | 354249 | 6244635 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 20.27 | 7.5 | 10.4 |
| GN-15 | GL-C | 355255 | 6247050 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 19.83 | 11.2 | 10.3 |
| GN-16 | GL-C | 356435 | 6248026 | 12-Sep-18 | 11.0 | 13-Sep-18 | 11.0 | 19.67 | 6.8 | 7.0 |
| GN-17 | BR-D | 338303 | 6245102 | 13-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 48.20 | 7.2 | 9.9 |
| GN-18 | GL-A | 349215 | 6244368 | 13-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 48.42 | 11.1 | 11.7 |
| GN-19 | GL-B | 353846 | 6244089 | 13-Sep-18 | 11.0 | 14-Sep-18 | 11.0 | 27.75 | 9.2 | 9.6 |
| GN-20 | GL-B | 354247 | 6244629 | 13-Sep-18 | 11.0 | 14-Sep-18 | 11.0 | 26.58 | 8.4 | 9.8 |

Table A1-2: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the future Keeyask reservoir, fall 2018 (continued).

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-21 | GL-C | 355229 | 6246908 | 13-Sep-18 | 11.0 | 14-Sep-18 | 10.5 | 25.58 | 8.5 | 11.4 |
| GN-22 | GL-C | 356388 | 6247998 | 13-Sep-18 | 11.0 | 14-Sep-18 | 10.5 | 24.42 | 6.9 | 8.4 |
| GN-23 | GL-C | 356609 | 6248022 | 14-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 19.42 | 6.0 | 7.4 |
| GN-24 | GL-C | 355236 | 6246973 | 14-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 22.50 | 9.5 | 8.6 |
| GN-25 | GL-B | 354367 | 6244636 | 14-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 21.25 | 11.1 | 12.2 |
| GN-26 | GL-B | 354040 | 6244477 | 14-Sep-18 | 11.0 | 15-Sep-18 | 10.5 | 20.33 | 12.8 | 9.4 |
| GN-27 | GL-A | 347839 | 6243704 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 24.33 | 10.4 | 9.8 |
| GN-28 | BR-D | 338340 | 6245148 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 22.50 | 9.1 | 7.8 |
| GN-29 | GL-A | 349102 | 6244392 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 22.92 | 10.1 | 10.7 |
| GN-30 | GL-B | 354218 | 6244495 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 23.17 | 11.2 | 8.9 |
| GN-31 | GL-B | 354594 | 6244678 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 23.08 | 10.4 | 12.0 |
| GN-32 | GL-C | 356871 | 6248090 | 15-Sep-18 | 10.5 | 16-Sep-18 | 10.0 | 23.08 | 7.7 | 10.8 |
| GN-33 | BR-D | 338331 | 6245130 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 23.58 | 9.1 | 9.9 |
| GN-34 | GL-A | 347942 | 6243690 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 23.25 | 7.3 | 12.6 |
| GN-35 | GL-A | 349026 | 6244355 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 23.33 | 9.5 | 8.7 |
| GN-36 | GL-B | 354312 | 6244769 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 24.08 | 7.6 | 8.3 |
| GN-37 | GL-B | 354581 | 6244473 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 23.08 | 10.0 | 10.9 |
| GN-38 | GL-C | 357044 | 6248087 | 16-Sep-18 | 10.0 | 17-Sep-18 | 9.5 | 23.58 | 9.1 | 7.9 |
| GN-39 | BR-D | 338326 | 6245121 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 24.00 | 10.2 | 9.7 |
| GN-40 | GL-A | 347954 | 6243697 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 23.83 | 7.7 | 10.4 |
| GN-41 | GL-A | 349044 | 6244362 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 23.75 | 10.3 | 12.0 |

Table A1-2: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in the future Keeyask reservoir, fall 2018 (continued).

| Site | Zone | UTM Location |  | Set Date | Set Water Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-42 | GL-B | 354598 | 6244499 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 23.17 | 9.1 | 10.5 |
| GN-43 | GL-B | 354312 | 6244781 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 23.08 | 7.0 | 8.3 |
| GN-44 | GL-C | 356386 | 6248021 | 17-Sep-18 | 9.5 | 18-Sep-18 | 9.0 | 23.25 | 7.3 | 8.7 |
| GN-45 | GL-A | 347965 | 6243677 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 22.92 | 5.6 | 9.4 |
| GN-46 | GL-A | 348875 | 6244335 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 22.83 | 9.6 | 9.3 |
| GN-47 | GL-B | 354710 | 6244550 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 22.83 | 10.7 | 8.1 |
| GN-48 | GL-B | 354425 | 6244811 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 22.83 | 10.5 | 11.3 |
| GN-49 | GL-C | 356068 | 6247803 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 23.33 | 6.5 | 7.1 |
| GN-50 | GL-C | 356426 | 6248046 | 18-Sep-18 | 9.0 | 19-Sep-18 | 9.0 | 23.00 | 7.5 | 7.9 |

Table A1-3: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in Stephens Lake, fall 2018.

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-01 | STL-A | 366622 | 6247496 | 9-Sep-18 | 12.0 | 11-Sep-18 | 12.0 | 41.33 | 16.8 | - |
| GN-02 | STL-A | 366842 | 6247513 | 9-Sep-18 | 12.0 | 11-Sep-18 | 12.0 | 40.50 | - | - |
| GN-03 | STL-A | 366706 | 6247405 | 9-Sep-18 | 12.0 | 11-Sep-18 | 12.0 | 41.58 | - | - |
| GN-04 | STL-B | 370233 | 6148460 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 23.92 | 17.0 | 15.6 |
| GN-05 | STL-B | 369310 | 6248795 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 23.83 | 17.1 | 16.2 |
| GN-06 | STL-B | 368918 | 6248711 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 24.00 | 17.4 | 14.4 |
| GN-07 | STL-A | 366686 | 6247648 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 24.50 | 14.3 | 15.3 |
| GN-08 | STL-A | 366622 | 6247506 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 24.67 | 15.8 | 14.6 |
| GN-09 | STL-A | 366538 | 6247395 | 11-Sep-18 | 12.0 | 12-Sep-18 | 11.5 | 24.33 | 15.0 | 16.9 |
| GN-10 | STL-B | 370059 | 6248422 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 27.17 | 18.0 | 17.0 |
| GN-11 | STL-B | 369357 | 6248807 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 25.32 | 17.0 | 14.0 |
| GN-12 | STL-B | 369071 | 6248742 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 24.92 | 14.0 | 15.5 |
| GN-13 | STL-A | 366392 | 6247554 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 21.92 | 14.6 | 14.1 |
| GN-14 | STL-A | 366613 | 6247499 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 22.05 | 15.4 | 14.4 |
| GN-14 | STL-A | 366613 | 6247499 | 13-Sep-18 | 11.0 | 14-Sep-18 | - | 28.83 | 15.4 | 14.4 |
| GN-15 | STL-B | 369063 | 6248843 | 12-Sep-18 | 11.5 | 13-Sep-18 | 11.0 | 21.92 | 14.8 | 18.3 |
| GN-16 | STL-A | 366138 | 6247103 | 13-Sep-18 | 11.0 | 14-Sep-18 | - | 30.08 | 9.2 | 14.0 |
| GN-17 | STL-B | 369134 | 6248729 | 13-Sep-18 | 11.0 | 15-Sep-18 | - | 46.42 | 14.9 | 14.8 |
| GN-18 | STL-B | 368953 | 6248710 | 13-Sep-18 | 11.0 | 15-Sep-18 | - | 45.17 | 16.0 | 15.1 |
| GN-18 | STL-B | 368953 | 6248710 | 15-Sep-18 | - | 16-Sep-18 | - | 25.50 | 16.0 | 15.1 |
| GN-19 | STL-B | 369232 | 6248308 | 13-Sep-18 | 11.0 | 15-Sep-18 | - | 45.58 | 16.9 | 18.0 |
| GN-20 | STL-B | 369405 | 6248099 | 13-Sep-18 | 11.0 | 15-Sep-18 | - | 45.50 | 15.0 | 14.0 |
| GN-21 | STL-A | 366639 | 6247288 | 14-Sep-18 | - | 15-Sep-18 | - | 20.33 | 12.9 | 16.9 |

Table A1-3: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in Stephens Lake, fall 2018 (continued).

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-21 | STL-A | 366639 | 6247288 | 15-Sep-18 | - | 16-Sep-18 | - | 23.92 | 12.9 | 16.9 |
| GN-22 | STL-A | 366675 | 6247251 | 14-Sep-18 | - | 15-Sep-18 | - | 21.00 | 13.3 | 18.0 |
| GN-23 | STL-B | 368994 | 6248690 | 15-Sep-18 | - | 16-Sep-18 | - | 25.25 | 16.1 | 14.5 |
| GN-24 | STL-B | 368775 | 6248602 | 15-Sep-18 | - | 16-Sep-18 | - | 22.33 | 16.2 | 17.9 |
| GN-25 | STL-B | 368772 | 6248527 | 15-Sep-18 | - | 16-Sep-18 | - | 22.25 | 15.2 | 17.8 |
| GN-26 | STL-A | 366649 | 6247339 | 15-Sep-18 | - | 16-Sep-18 | - | 23.00 | 18.9 | 16.9 |
| GN-27 | STL-B | 368890 | 6248610 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 23.97 | 17.1 | 16.3 |
| GN-28 | STL-B | 368726 | 6248499 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 23.00 | 13.9 | 17.1 |
| GN-29 | STL-B | 368966 | 6248718 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 22.83 | - | - |
| GN-30 | STL-B | 369005 | 6248535 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 23.00 | - | - |
| GN-31 | STL-A | 366646 | 6247305 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 23.55 | - | - |
| GN-32 | STL-A | 366669 | 6247561 | 16-Sep-18 | - | 17-Sep-18 | 10.0 | 22.50 | - | - |
| GN-33 | STL-B | 368729 | 6248385 | 17-Sep-18 | 10.0 | 18-Sep-18 | - | 25.25 | 9.6 | 14.7 |
| GN-34 | STL-B | 368893 | 6248647 | 17-Sep-18 | 10.0 | 18-Sep-18 | - | 25.92 | 17.3 | 16.4 |
| GN-35 | STL-B | 368064 | 6248639 | 17-Sep-18 | 10.0 | 18-Sep-18 | - | 23.38 | 15.2 | 15.0 |
| GN-36 | STL-B | 369044 | 6248404 | 17-Sep-18 | 10.0 | 18-Sep-18 | - | 24.25 | 14.4 | 13.1 |
| GN-37 | STL-A | 367725 | 6248105 | 17-Sep-18 | 10.0 | 18-Sep-18 | - | 21.22 | 15.0 | 15.5 |
| GN-38 | STL-A | 366654 | 6247334 | 17-Sep-18 | 10 | 18-Sep-18 | - | 23.92 | 18.4 | 17.4 |
| GN-39 | STL-B | 367900 | 6248148 | 18-Sep-18 | - | 19-Sep-18 | - | 27.00 | 14.8 | 14.5 |
| GN-40 | STL-B | 367899 | 6248228 | 18-Sep-18 | - | 19-Sep-18 | - | 26.67 | 14.4 | 15.6 |
| GN-41 | STL-B | 368895 | 6248434 | 18-Sep-18 | - | 19-Sep-18 | - | 24.33 | 16.0 | 16.6 |
| GN-41 | STL-B | 368895 | 6248434 | 19-Sep-18 | - | 20-Sep-18 | - | 22.08 | 16.0 | 16.6 |

Table A1-3: Location and site-specific physical measurements collected at gillnetting sites during juvenile Lake Sturgeon investigations conducted in Stephens Lake, fall 2018 (continued).

| Site | Zone | UTM Location |  | Set Date | Set Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Pull Date | Pull Water <br> Temp ( ${ }^{\circ} \mathrm{C}$ ) | Duration (dec.hrs) | Water Depth (m) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Easting | Northing |  |  |  |  |  | Start | End |
| GN-42 | STL-B | 369724 | 6248541 | 18-Sep-18 | - | 19-Sep-18 | - | 23.05 | 20.9 | 19.6 |
| GN-43 | STL-B | 368245 | 6247971 | 18-Sep-18 | - | 19-Sep-18 | - | 23.95 | 15.9 | 14.7 |
| GN-44 | STL-A | 367401 | 6247789 | 18-Sep-18 | - | 19-Sep-18 | - | 25.13 | 15.5 | 14.4 |
| GN-45 | STL-B | 367802 | 6248028 | 19-Sep-18 | - | 20-Sep-18 | - | 23.42 | 15.2 | 16.2 |
| GN-46 | STL-B | 368243 | 6247889 | 19-Sep-18 | - | 20-Sep-18 | - | 21.50 | 15.9 | 15.9 |
| GN-47 | STL-B | 368889 | 6248443 | 19-Sep-18 | - | 20-Sep-18 | - | 20.30 | 16.6 | 16.0 |
| GN-48 | STL-B | 368125 | 6248194 | 19-Sep-18 | - | 20-Sep-18 | - | 20.87 | 16.4 | 13.4 |
| GN-49 | STL-B | 367971 | 6247732 | 19-Sep-18 | - | 20-Sep-18 | - | 20.75 | 15.8 | 17.1 |

## APPENDIX 2: <br> BIOLOGICAL AND TAG INFORMATION FOR LAKE STURGEON CAPTURED IN FALL 2018.

Table A2-1: Biological and tag information for Lake Sturgeon captured in the UpperSplit Lake Area (Burntwood River and Split Lake), fall 2018.86
Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018 ..... 89
Table A2-3: Biological and tag information for Lake Sturgeon captured in Stephens Lake, fall 2018. ..... 94

Table A2-1: Biological and tag information for Lake Sturgeon captured in the Upper Split Lake Area (Burntwood River and Split Lake), fall 2018.

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113526 | 900226000153713 | 641 | 723 | 1750 | 5 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113528 | 900226000153755 | 660 | 747 | 2200 | 12 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113529 | 900226000153756 | 660 | 751 | 2050 | 13 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113530 | 900226000153738 | 509 | 583 | 850 | 5 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113531 | 900226000153777 | 400 | 455 | 450 | 5 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113532 | 900226000153789 | 490 | 551 | 800 | 5 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113533 | 9001380344522 | 323 | 366 | 250 | 2 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | 113534 | 900226000153700 | 500 | 569 | 900 | 5 |
| Split Lake | GN-01 | SPL-A | 11-Sep-18 | - | - | 520 | 587 | 1000 | 5 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113535 | 900226000153766 | 534 | 598 | 950 | 4 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113536 | 900226000153705 | 616 | 692 | 1600 | 7 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113537 | 900226000153792 | 841 | 931 | 4650 | - |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113538 | 900226000153780 | 730 | 813 | 3650 | 8 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113539 | 900226000153706 | 689 | 762 | 2550 | 8 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113540 | 900226000153754 | 522 | 695 | 1100 | 5 |
| Split Lake | GN-02 | SPL-A | 11-Sep-18 | 113541 | 900226000153701 | 441 | 486 | 675 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113543 | 900226000153786 | 528 | 601 | 1100 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113544 | 900226000153776 | 538 | 605 | 1200 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113545 | 9001380344115 | 302 | 341 | 175 | 3 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113546 | 900226000153742 | 363 | 522 | 700 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113547 | 900226000153710 | 515 | 580 | 1100 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113548 | 900226000153793 | 562 | 640 | 1300 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113549 | 900226000153795 | 496 | 564 | 750 | 4 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113501 | 900226000153729 | 447 | 509 | 700 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113502 | 900226000153707 | 426 | 489 | 650 | 5 |
| Split Lake | GN-03 | SPL-A | 11-Sep-18 | 113503 | 900067000121368 | 335 | 375 | 250 | 2 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | - | - | 676 | 762 | 2450 | 7 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | - | - | 497 | 552 | 1000 | 5 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | - | - | 540 | 614 | 1200 | 5 |

Table A2-1: Biological and tag information for Lake Sturgeon captured in the Upper Split Lake Area (Burntwood River and Split Lake), fall 2018 (continued).

| Waterbody | Site | Zone | Date | $\begin{gathered} \text { Floy-tag } \\ \# \end{gathered}$ | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113504 | 900226000153733 | 671 | 741 | 2250 | 7 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113505 | 900226000153751 | 511 | 585 | 1200 | 5 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113506 | 900226000153743 | 505 | 572 | 900 | 5 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113507 | 900226000153762 | 430 | 494 | 550 | 5 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113508 | 900226000153721 | 706 | 798 | 2500 | 7 |
| Split Lake | GN-04 | SPL-A | 11-Sep-18 | 113509 | 900226000768511 | 541 | 613 | 1250 | 5 |
| Split Lake | GN-08 | SPL-A | 12-Sep-18 | 113510 | 900226000153796 | 666 | 741 | 2500 | 7 |
| Split Lake | GN-07 | SPL-A | 12-Sep-18 | 113511 | 900226000153718 | 745 | 834 | 3450 | 7 |
| Split Lake | GN-07 | SPL-A | 12-Sep-18 | 113512 | 900226000153726 | 570 | 633 | 1250 | 5 |
| Split Lake | GN-07 | SPL-A | 12-Sep-18 | 113513 | 900226000153785 | 712 | 806 | 2600 | 7 |
| Split Lake | GN-07 | SPL-A | 12-Sep-18 | 113514 | 900226000768550 | 770 | 871 | 4150 | 12 |
| Split Lake | GN-07 | SPL-A | 12-Sep-18 | 79284 | 900226000153736 | 683 | 779 | 2200 | 7 |
| Split Lake | GN-06 | SPL-A | 12-Sep-18 | 113516 | 900226000768523 | 732 | 830 | 2250 | 10 |
| Split Lake | GN-06 | SPL-A | 12-Sep-18 | 113517 | 900226000768598 | 560 | 632 | 1375 | 6 |
| Split Lake | GN-05 | SPL-A | 12-Sep-18 | 113519 | 900226000153759 | 685 | 753 | 2450 | 11 |
| Split Lake | GN-05 | SPL-A | 12-Sep-18 | 91364 | 900226000548175 | 996 | 1160 | 7350 | - |
| Split Lake | GN-05 | SPL-A | 12-Sep-18 | 113520 | 900226000153709 | 805 | 912 | 4250 | - |
| Split Lake | GN-05 | SPL-A | 12-Sep-18 | 91600 | 900226000153775 | 796 | 892 | 4050 | 13 |
| Split Lake | GN-05 | SPL-A | 12-Sep-18 | 113521 | 900226000153763 | 814 | 916 | 4150 | - |
| Burntwood River | GN-09 | BWR-A | 13-Sep-18 | 88179 | 900226000893378 | 749 | 822 | 2650 | - |
| Burntwood River | GN-09 | BWR-A | 13-Sep-18 | 113522 | 900226000767451 | 636 | 712 | 1650 | 19 |
| Burntwood River | GN-09 | BWR-A | 13-Sep-18 | 113523 | 900226000153771 | 492 | 563 | 900 | 8 |
| Burntwood River | GN-10 | BWR-A | 13-Sep-18 | 113525 | 900226000153715 | 764 | 863 | 3000 | 16 |
| Burntwood River | GN-12 | BWR-A | 13-Sep-18 | 113725 | 900226000153725 | 367 | 415 | 350 | 3 |
| Burntwood River | GN-13 | BWR-A | 14-Sep-18 | 113724 | 900226000153757 | 491 | 561 | 900 | 8 |
| Burntwood River | GN-16 | BWR-B | 14-Sep-18 | 113723 | 972273000041195 | 205 | 228 | 25 | 1 |
| Burntwood River | GN-16 | BWR-B | 14-Sep-18 | 113722 | 900067000121588 | 351 | 401 | 225 | 3 |
| Burntwood River | GN-22 | BWR-C | 16-Sep-18 | 113721 | 972273000041200 | 278 | 316 | 100 | 2 |
| Burntwood River | GN-26 | BWR-C | 17-Sep-18 | 113719 | 900067000121317 | 272 | 308 | 175 | 2 |
| Burntwood River | GN-27 | BWR-C | 17-Sep-18 | 113718 | 900226000153773 | 403 | 456 | 475 | 5 |

Aquatic Effects Monitoring Plan

Table A2-1: Biological and tag information for Lake Sturgeon captured in the Upper Split Lake Area (Burntwood River and Split Lake), fall 2018 (continued).

| Waterbody | Site | Zone | Date | Floy-tag <br> \# | Pit-tag \# | Fork Length <br> $\mathbf{( m m )}$ | Total <br> Length <br> $\mathbf{( m m )}$ | Weight <br> $(\mathbf{g})$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Split Lake | GN-28 | SPL-A | 17-Sep-18 | 113717 | 900226000767450 | 796 | 882 | 3400 |
| Age |  |  |  |  |  |  |  |  |

Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018.

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | GN-01 | GL-C | 10-Sep-18 | 113001 | 900226000327580 | 725 | 834 | 2950 | 11 |
| Future Keeyask Reservoir | GN-02 | GL-C | 10-Sep-18 | - | 900067000121507 | 296 | 336 | 200 | 2 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113002 | 900226000327570 | 430 | 486 | 550 | 4 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113003 | 900226000327527 | 429 | 494 | 500 | 4 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113004 | 900226000327530 | 381 | 441 | 400 | 4 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113005 | 900226000327508 | 354 | 401 | 300 | 3 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113006 | 900067000112175 | 353 | 409 | 300 | 2 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113007 | 900226000327539 | 320 | 323 | 250 | 2 |
| Future Keeyask Reservoir | GN-06 | GL-B | 11-Sep-18 | 113008 | 900067000121578 | 286 | 322 | 200 | 2 |
| Future Keeyask Reservoir | GN-05 | GL-C | 11-Sep-18 | 113009 | 900226000327556 | 786 | 876 | 3450 | 10 |
| Future Keeyask Reservoir | GN-04 | GL-C | 11-Sep-18 | 103649 | 900226000629219 | 873 | 984 | 5300 | - |
| Future Keeyask Reservoir | GN-04 | GL-C | 11-Sep-18 | 106462 | 900226000893868 | 707 | 797 | 2450 | 9 |
| Future Keeyask Reservoir | GN-04 | GL-C | 11-Sep-18 | 113010 | 900226000327566 | 715 | 799 | 3000 | 10 |
| Future Keeyask Reservoir | GN-04 | GL-C | 11-Sep-18 | 113011 | 900226000327555 | 764 | 853 | 3700 | 10 |
| Future Keeyask Reservoir | GN-04 | GL-C | 11-Sep-18 | 113012 | 900226000327543 | 714 | 815 | 3000 | 11 |
| Future Keeyask Reservoir | GN-08 | GL-B | 12-Sep-18 | 113013 | 900226000327504 | 716 | 802 | 2500 | 10 |
| Future Keeyask Reservoir | GN-08 | GL-B | 12-Sep-18 | 113014 | 900226000327572 | 354 | 413 | 400 | 3 |
| Future Keeyask Reservoir | GN-08 | GL-B | 12-Sep-18 | 107115 | 900226000768485 | 955 | 1084 | 7100 | - |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | 113015 | 900226000327577 | 567 | 648 | 1350 | 8 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | 113016 | 900226000327568 | 357 | 395 | 400 | 3 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | 900067000121326 | 285 | 327 | 300 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | 113017 | 900226000327581 | 315 | 350 | 300 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | 113018 | 900226000327584 | 323 | 358 | 300 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | 113019 | 900226000327533 | 307 | 350 | 300 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | 900067000121396 | 293 | 331 | 150 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | 972273000041199 | 282 | 330 | 150 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | 900043000103752 | 282 | 309 | 150 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | 900067000121633 | 269 | 296 | 100 | 2 |
| Future Keeyask Reservoir | GN-09 | GL-C | 12-Sep-18 | - | - | 300 | 338 | 200 | 2 |
| Future Keeyask Reservoir | GN-10 | GL-C | 12-Sep-18 | 103116 | 900226000768467 | 767 | 871 | 3200 | 9 |
| Future Keeyask Reservoir | GN-10 | GL-C | 12-Sep-18 | 89658 | 900226000327546 | 770 | 858 | 3100 | 10 |
| Future Keeyask Reservoir | GN-10 | GL-C | 12-Sep-18 | 113020 | 900226000327535 | 752 | 874 | 3200 | 10 |

Aquatic Effects Monitoring Plan

Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018 (continued).

| Waterbody | Site | Zone | Date | Floy-tag <br> $\#$ | Pit-tag \# | Fork Length <br> (mm) | Total <br> Length <br> (mm) | Weight <br> (g) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :---: | :---: | :---: |
| Age |  |  |  |  |  |  |  |  |

Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018 (continued).

| Waterbody | Site | Zone | Date | $\begin{aligned} & \text { Floy-tag } \\ & \quad \# \end{aligned}$ | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | GN-22 | GL-C | 14-Sep-18 | - | 900067000121390 | 214 | 241 | 50 | 1 |
| Future Keeyask Reservoir | GN-21 | GL-C | 14-Sep-18 | 113046 | 900226000327518 | 764 | 886 | 3300 | 10 |
| Future Keeyask Reservoir | GN-20 | GL-B | 14-Sep-18 | 113047 | 900226000327520 | 463 | 539 | 650 | 5 |
| Future Keeyask Reservoir | GN-20 | GL-B | 14-Sep-18 | 113048 | 900067000112161 | 403 | 458 | 350 | 2 |
| Future Keeyask Reservoir | GN-20 | GL-B | 14-Sep-18 | 113049 | 900226000327559 | 314 | 348 | 100 | 2 |
| Future Keeyask Reservoir | GN-19 | GL-B | 14-Sep-18 | 113050 | 900226000327554 | 699 | 805 | 2700 | 12 |
| Future Keeyask Reservoir | GN-19 | GL-B | 14-Sep-18 | 113005 | 900226000327537 | 692 | 787 | 2050 | 10 |
| Future Keeyask Reservoir | GN-23 | GL-C | 15-Sep-18 | 113825 | 900226000767269 | 745 | 842 | 2950 | 10 |
| Future Keeyask Reservoir | GN-23 | GL-C | 15-Sep-18 | 113824 | 900226000768413 | 708 | 801 | 2450 | 10 |
| Future Keeyask Reservoir | GN-17 | BR-D | 15-Sep-18 | 113823 | 900226000327586 | 842 | 956 | 4550 | - |
| Future Keeyask Reservoir | GN-17 | BR-D | 15-Sep-18 | 113822 | 900226000767273 | 758 | 862 | 3450 | 10 |
| Future Keeyask Reservoir | GN-18 | GL-A | 15-Sep-18 | 113821 | 900043000103674 | 486 | 549 | 650 | 5 |
| Future Keeyask Reservoir | GN-26 | GL-B | 15-Sep-18 | 113820 | 900226000767281 | 305 | 354 | 200 | 2 |
| Future Keeyask Reservoir | GN-25 | GL-B | 15-Sep-18 | 113819 | 900226000327549 | 569 | 649 | 1150 | 7 |
| Future Keeyask Reservoir | GN-25 | GL-B | 15-Sep-18 | 113818 | 900226000767262 | 351 | 386 | 250 | 3 |
| Future Keeyask Reservoir | GN-25 | GL-B | 15-Sep-18 | 113817 | 900226000327598 | 336 | 386 | 200 | 2 |
| Future Keeyask Reservoir | GN-25 | GL-B | 15-Sep-18 | 113816 | 900226000327573 | 325 | 374 | 200 | 2 |
| Future Keeyask Reservoir | GN-25 | GL-B | 15-Sep-18 | 113815 | 900226000767201 | 310 | 346 | 150 | 2 |
| Future Keeyask Reservoir | GN-24 | GL-C | 15-Sep-18 | 106463 | 900226000154225 | 723 | 824 | 2850 | 10 |
| Future Keeyask Reservoir | GN-24 | GL-C | 15-Sep-18 | 113814 | 900226000629005 | 675 | 764 | 2450 | 9 |
| Future Keeyask Reservoir | GN-28 | BR-D | 16-Sep-18 | 113813 | 900226000327557 | 759 | 850 | 3200 | 10 |
| Future Keeyask Reservoir | GN-28 | BR-D | 16-Sep-18 | 113812 | 900226000327547 | 655 | 739 | 1900 | 10 |
| Future Keeyask Reservoir | GN-28 | BR-D | 16-Sep-18 | 113811 | 900226000767255 | 579 | 665 | 1450 | 6 |
| Future Keeyask Reservoir | GN-27 | GL-A | 16-Sep-18 | 113810 | 900226000327587 | 727 | 830 | 2750 | 10 |
| Future Keeyask Reservoir | GN-27 | GL-A | 16-Sep-18 | 113809 | 900226000327590 | 655 | 739 | 2150 | 10 |
| Future Keeyask Reservoir | GN-29 | GL-A | 16-Sep-18 | 112543 | 900226000629452 | 574 | 651 | 1100 | $7^{\text {a }}$ |
| Future Keeyask Reservoir | GN-29 | GL-A | 16-Sep-18 | 113808 | 900226000767293 | 504 | 574 | 850 | 5 |
| Future Keeyask Reservoir | GN-30 | GL-B | 16-Sep-18 | 113807 | 900226000767288 | 742 | 836 | 2950 | 10 |
| Future Keeyask Reservoir | GN-32 | GL-C | 16-Sep-18 | 113806 | 900226000327571 | 519 | 595 | 1000 | 5 |
| Future Keeyask Reservoir | GN-32 | GL-C | 16-Sep-18 | 113805 | 900226000327562 | 375 | 475 | 350 | 3 |
| Future Keeyask Reservoir | GN-32 | GL-C | 16-Sep-18 | 113804 | 900226000327519 | 316 | 353 | 200 | 2 |
| Future Keeyask Reservoir | GN-32 | GL-C | 16-Sep-18 | - | 972273000041186 | 219 | 250 | 100 | 1 |

Aquatic Effects Monitoring Plan

Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018 (continued).

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | GN-32 | GL-C | 16-Sep-18 | - | 900067000121601 | 212 | 240 | 100 | 1 |
| Future Keeyask Reservoir | GN-34 | GL-A | 17-Sep-18 | 113803 | 900226000327526 | 943 | 1030 | 7650 | - |
| Future Keeyask Reservoir | GN-34 | GL-A | 17-Sep-18 | 113802 | 900226000327538 | 361 | 414 | 350 | 3 |
| Future Keeyask Reservoir | GN-34 | GL-A | 17-Sep-18 | - | 900043000103776 | 201 | 225 | 100 | 1 |
| Future Keeyask Reservoir | GN-35 | GL-A | 17-Sep-18 | 113801 | 900226000327525 | 630 | 724 | 1850 | 10 |
| Future Keeyask Reservoir | GN-35 | GL-A | 17-Sep-18 | 113826 | 900226000327591 | 304 | 345 | 200 | 2 |
| Future Keeyask Reservoir | GN-35 | GL-A | 17-Sep-18 | - | 900067000121534 | 280 | 312 | 150 | 2 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113827 | 900226000327541 | 741 | 830 | 2700 | 12 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113828 | 900226000327531 | 701 | 793 | 2600 | 10 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113829 | 900226000327513 | 712 | 795 | 2350 | 10 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113830 | 900226000327550 | 670 | 761 | 2150 | 10 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113831 | 900067000055689 | 440 | 511 | 500 | 4 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113832 | 900067000055259 | 405 | 466 | 400 | 4 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | 113833 | 900067000112140 | 373 | 428 | 300 | 2 |
| Future Keeyask Reservoir | GN-37 | GL-B | 17-Sep-18 | - | 900067000121476 | 271 | 309 | 150 | 2 |
| Future Keeyask Reservoir | GN-36 | GL-B | 17-Sep-18 | 105691 | 900043000103824 | 668 | 766 | 1850 | $10^{\text {a }}$ |
| Future Keeyask Reservoir | GN-36 | GL-B | 17-Sep-18 | 113834 | 900226000767217 | 453 | 512 | 500 | 4 |
| Future Keeyask Reservoir | GN-36 | GL-B | 17-Sep-18 | 113835 | 900067000058508 | 453 | 519 | 500 | 4 |
| Future Keeyask Reservoir | GN-36 | GL-B | 17-Sep-18 | 113836 | 900067000112127 | 400 | 463 | 400 | 2 |
| Future Keeyask Reservoir | GN-36 | GL-B | 17-Sep-18 | 113837 | 900226000327521 | 340 | 382 | 250 | 3 |
| Future Keeyask Reservoir | GN-38 | GL-C | 17-Sep-18 | 112502 | 900226000893677 | 669 | 765 | 2300 | 10 |
| Future Keeyask Reservoir | GN-40 | GL-A | 18-Sep-18 | 113838 | 900226000327558 | 840 | 968 | 4050 | - |
| Future Keeyask Reservoir | GN-40 | GL-A | 18-Sep-18 | 113839 | 900067000059354 | 318 | 365 | 200 | 2 |
| Future Keeyask Reservoir | GN-42 | GL-B | 18-Sep-18 | 113840 | 900226000327536 | 730 | 821 | 2500 | 10 |
| Future Keeyask Reservoir | GN-42 | GL-B | 18-Sep-18 | - | 900067000121312 | 285 | 325 | 200 | 2 |
| Future Keeyask Reservoir | GN-43 | GL-B | 18-Sep-18 | 113841 | 900043000103668 | 489 | 556 | 700 | 5 |
| Future Keeyask Reservoir | GN-43 | GL-B | 18-Sep-18 | 113842 | 900226000327522 | 303 | 340 | 200 | 2 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113843 | 900226000327589 | 569 | 658 | 1300 | 6 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113844 | 900226000327514 | 534 | 606 | 1150 | 6 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113845 | 900226000327585 | 477 | 558 | 800 | 5 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113846 | 900226000327592 | 566 | 646 | 1200 | 7 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113847 | 900226000767260 | 435 | 506 | 500 | 4 |

Table A2-2: Biological and tag information for Lake Sturgeon captured in the future Keeyask reservoir, fall 2018 (continued).

| Waterbody | Site | Zone | Date | $\begin{aligned} & \text { Floy-tag } \\ & \quad \# \end{aligned}$ | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113848 | 900067000056730 | 381 | 438 | 400 | 2 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113849 | 900226000327502 | 325 | 370 | 250 | 2 |
| Future Keeyask Reservoir | GN-44 | GL-C | 18-Sep-18 | 113850 | 900226000327507 | 305 | 345 | 200 | 2 |
| Future Keeyask Reservoir | GN-46 | GL-A | 19-Sep-18 | 113151 | 900226000767272 | 331 | 376 | 200 | 2 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113152 | 900226000327578 | 633 | 724 | 1700 | 10 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113153 | 900067000058648 | 479 | 545 | 700 | 4 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113154 | 900067000055405 | 462 | 535 | 600 | 4 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113155 | 900226000327540 | 419 | 473 | 550 | 4 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113156 | 900226000327561 | 302 | 345 | 250 | 2 |
| Future Keeyask Reservoir | GN-47 | GL-B | 19-Sep-18 | 113157 | 900226000327534 | 310 | 353 | 300 | 2 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113158 | 900226000327553 | 688 | 805 | 2350 | 10 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113159 | 900226000327529 | 584 | 657 | 1200 | 7 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113160 | 900226000327528 | 479 | 545 | 750 | 6 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113161 | 900067000055447 | 448 | 518 | 500 | 4 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113162 | 900226000327544 | 304 | 345 | 200 | 2 |
| Future Keeyask Reservoir | GN-48 | GL-B | 19-Sep-18 | 113163 | 900226000327512 | 310 | 352 | 200 | 2 |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | 91383 | 900226000629177 | 1031 | 1112 | 8500 | - |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | 113164 | 900226000327523 | 654 | 747 | 1800 | 8 |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | 113165 | 900226000327551 | 590 | 670 | 1550 | 7 |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | 113166 | 900226000327597 | 524 | 591 | 900 | 5 |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | 113167 | 900226000327532 | 387 | 439 | 400 | 3 |
| Future Keeyask Reservoir | GN-50 | GL-C | 19-Sep-18 | - | 900067000110427 | 255 | 296 | 100 | 2 |

[^1]Table A2-3: Biological and tag information for Lake Sturgeon captured in Stephens Lake, fall 2018.

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | GN-02 | STL-A | 11-Sep-18 | - | - | 750 | 840 | 3450 | - |
| Stephens Lake | GN-01 | STL-A | 11-Sep-18 | 100673 | 900067000055481 | 455 | 525 | 750 | 4 |
| Stephens Lake | GN-01 | STL-A | 11-Sep-18 | 115144 | 900067000058474 | 490 | 562 | 875 | 4 |
| Stephens Lake | GN-03 | STL-A | 11-Sep-18 | 115145 | 900226000327820 | 574 | 655 | 1700 | 7 |
| Stephens Lake | GN-05 | STL-B | 12-Sep-18 | 115146 | 900226000893322 | 330 | 384 | 200 | 2 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 115147 | 900226000327949 | 430 | 500 | 500 | 3 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 115149 | 900226000154278 | 436 | 501 | 625 | 3 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 115150 | - | 394 | 451 | 375 | 3 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 115148 | 900226000327876 | 300 | 337 | 200 | 2 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 113251 | 900226000327847 | 635 | 724 | 1925 | 7 |
| Stephens Lake | GN-06 | STL-B | 12-Sep-18 | 101500 | 900226000628170 | 812 | 909 | 3650 | - |
| Stephens Lake | GN-07 | STL-A | 12-Sep-18 | 110580 | 900067000121406 | 386 | 435 | 300 | 3 |
| Stephens Lake | GN-07 | STL-A | 12-Sep-18 | - | 900067000121295 | 222 | 255 | 50 | 1 |
| Stephens Lake | GN-08 | STL-A | 12-Sep-18 | - | 900067000121303 | 241 | 275 | 75 | 1 |
| Stephens Lake | GN-14 | STL-A | 13-Sep-18 | 115826 | 900226000154078 | 837 | 945 | 4300 | - |
| Stephens Lake | GN-14 | STL-A | 13-Sep-18 | 113252 | 900226000327941 | 418 | 472 | 575 | 3 |
| Stephens Lake | GN-14 | STL-A | 13-Sep-18 | - | 972273000041181 | 307 | 344 | 225 | 2 |
| Stephens Lake | GN-11 | STL-B | 13-Sep-18 | 113253 | 900226000327855 | 765 | 825 | 3575 | 10 |
| Stephens Lake | GN-11 | STL-B | 13-Sep-18 | 113255 | 900226000327821 | 820 | 937 | 4925 | - |
| Stephens Lake | GN-12 | STL-B | 13-Sep-18 | 113256 | 900226000327812 | 435 | 496 | 550 | 3 |
| Stephens Lake | GN-12 | STL-B | 13-Sep-18 | 113257 | 900226000327851 | 505 | 590 | 975 | 5 |
| Stephens Lake | GN-18 | STL-B | 15-Sep-18 | 113258 | 900226000327993 | 410 | 466 | 550 | 3 |
| Stephens Lake | GN-18 | STL-B | 15-Sep-18 | 110994 | 900226000548945 | 522 | 595 | 1100 | 5 |
| Stephens Lake | GN-18 | STL-B | 15-Sep-18 | 113259 | 900067000055178 | 503 | 570 | 900 | 4 |
| Stephens Lake | GN-18 | STL-B | 15-Sep-18 | 112945 | 900226000893689 | 545 | 630 | 1150 | 5 |
| Stephens Lake | GN-18 | STL-B | 15-Sep-18 | 113260 | 900226000327828 | 421 | 480 | 575 | 3 |
| Stephens Lake | GN-20 | STL-B | 15-Sep-18 | 92093 | - | 770 | 874 | 3300 | $10^{\text {a }}$ |
| Stephens Lake | GN-21 | STL-A | 15-Sep-18 | 110582 | 900067000055566 | 530 | 616 | 1050 | 5 |
| Stephens Lake | GN-21 | STL-A | 15-Sep-18 | 113261 | 900067000055526 | 495 | 566 | 850 | 4 |
| Stephens Lake | GN-21 | STL-A | 15-Sep-18 | 113263 | 900067000055507 | 497 | 566 | 875 | 4 |

Aquatic Effects Monitoring Plan

Table A2-3: Biological and tag information for Lake Sturgeon captured in Stephens Lake, fall 2018 (continued).

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | GN-21 | STL-A | 15-Sep-18 | 113264 | 900067000055170 | 480 | 550 | 775 | 4 |
| Stephens Lake | GN-21 | STL-A | 15-Sep-18 | 113265 | 900067000111882 | 354 | 410 | 375 | 2 |
| Stephens Lake | GN-24 | STL-B | 16-Sep-18 | 113266 | 900226000327809 | 421 | 480 | 525 | 4 |
| Stephens Lake | GN-25 | STL-B | 16-Sep-18 | 113267 | 900226000327943 | 500 | 575 | 825 | 4 |
| Stephens Lake | GN-18 | STL-B | 16-Sep-18 | 113268 | 900226000767617 | 392 | 446 | 475 | 3 |
| Stephens Lake | GN-23 | STL-B | 16-Sep-18 | 96513 | 900226000154249 | 531 | 602 | 1075 | $6^{\text {a }}$ |
| Stephens Lake | GN-28 | STL-B | 17-Sep-18 | 113269 | 900067000055076 | 502 | 568 | 700 | 4 |
| Stephens Lake | GN-28 | STL-B | 17-Sep-18 | 113270 | 900226000327892 | 498 | 564 | 975 | 5 |
| Stephens Lake | GN-27 | STL-B | 17-Sep-18 | 113271 | 900067000111990 | 375 | 416 | 350 | 2 |
| Stephens Lake | GN-27 | STL-B | 17-Sep-18 | 113272 | 900226000767664 | 406 | 450 | 525 | 3 |
| Stephens Lake | GN-27 | STL-B | 17-Sep-18 | 113274 | 900226000327845 | 335 | 378 | 275 | 2 |
| Stephens Lake | GN-30 | STL-B | 17-Sep-18 | 113275 | 900226000327926 | 765 | 874 | 3325 | 10 |
| Stephens Lake | GN-30 | STL-B | 17-Sep-18 | 113273 | 900226000327963 | 500 | 564 | 900 | 5 |
| Stephens Lake | GN-31 | STL-A | 17-Sep-18 | 113276 | 900226000327907 | 550 | 610 | 1050 | 5 |
| Stephens Lake | GN-37 | STL-A | 18-Sep-18 | 113277 | 900226000327966 | 402 | 451 | 400 | 3 |
| Stephens Lake | GN-37 | STL-A | 18-Sep-18 | - | 900043000103587 | 251 | 283 | 75 | 1 |
| Stephens Lake | GN-33 | STL-B | 18-Sep-18 | 111064 | 900067000121333 | 385 | 430 | 375 | 3 |
| Stephens Lake | GN-33 | STL-B | 18-Sep-18 | 113278 | 900226000327947 | 652 | 746 | 2050 | 7 |
| Stephens Lake | GN-33 | STL-B | 18-Sep-18 | 113279 | 900226000767677 | 435 | 492 | 525 | 3 |
| Stephens Lake | GN-34 | STL-B | 18-Sep-18 | - | 900067000121203 | 322 | 367 | 175 | 3 |
| Stephens Lake | GN-41 | STL-B | 19-Sep-18 | 113280 | 900226000327915 | 371 | 414 | 375 | 3 |
| Stephens Lake | GN-41 | STL-B | 19-Sep-18 | 113281 | 900226000327919 | 371 | 414 | 400 | 3 |
| Stephens Lake | GN-41 | STL-B | 19-Sep-18 | 113282 | 900226000327958 | 409 | 467 | 425 | 3 |
| Stephens Lake | GN-41 | STL-B | 19-Sep-18 | 113284 | 900226000327852 | 413 | 480 | 500 | 3 |
| Stephens Lake | GN-43 | STL-B | 19-Sep-18 | 113285 | 900067000113506 | 392 | 405 | 425 | 2 |
| Stephens Lake | GN-39 | STL-B | 19-Sep-18 | 113286 | 900226000327877 | 420 | 479 | 525 | 3 |
| Stephens Lake | GN-40 | STL-B | 19-Sep-18 | 101994 | 9001380344626 | 430 | 477 | 600 | $4^{\text {a }}$ |
| Stephens Lake | GN-40 | STL-B | 19-Sep-18 | 113287 | 900226000327881 | 590 | 681 | 1500 | 7 |
| Stephens Lake | GN-40 | STL-B | 19-Sep-18 | 113288 | 900226000153878 | 366 | 412 | 325 | 3 |
| Stephens Lake | GN-44 | STL-A | 19-Sep-18 | 113289 | 900226000327937 | 540 | 609 | 1300 | 5 |
| Stephens Lake | GN-47 | STL-B | 20-Sep-18 | 113290 | 900226000327902 | 784 | 890 | 4325 | 10 |

Table A2-3: Biological and tag information for Lake Sturgeon captured in Stephens Lake, fall 2018 (continued).

| Waterbody | Site | Zone | Date | Floy-tag \# | Pit-tag \# | Fork Length (mm) | Total Length (mm) | Weight (g) | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | GN-41 | STL-B | 20-Sep-18 | 112924 | 900226000768894 | 410 | 447 | 550 | 3 |
| Stephens Lake | GN-41 | STL-B | 20-Sep-18 | 113291 | 900226000327980 | 452 | 505 | 650 | 5 |
| Stephens Lake | GN-48 | STL-B | 20-Sep-18 | 113292 | 900226000327973 | 522 | 594 | 1125 | 5 |
| Stephens Lake | GN-48 | STL-B | 20-Sep-18 | 113293 | 900067000058406 | 470 | 535 | 700 | 4 |
| Stephens Lake | GN-48 | STL-B | 20-Sep-18 | 113294 | 900226000327884 | 391 | 436 | 475 | 3 |
| Stephens Lake | GN-48 | STL-B | 20-Sep-18 | 113295 | 900067000059240 | 350 | 402 | 300 | 2 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 100674 | 900067000112046 | 357 | 411 | 300 | 2 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 113296 | 900067000055532 | 472 | 532 | 675 | 4 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 113297 | 900067000112569 | 346 | 396 | 275 | 2 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 113298 | 900226000327925 | 410 | 468 | 525 | 3 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 113299 | 900226000327930 | 375 | 438 | 400 | 3 |
| Stephens Lake | GN-45 | STL-B | 20-Sep-18 | 113300 | 900067000112320 | 346 | 403 | 300 | 2 |
| Stephens Lake | GN-49 | STL-B | 20-Sep-18 | 113283 | 900067000055210 | 464 | 532 | 700 | 4 |

a - Ages assigned based on structures aged in a previous study year.

## APPENDIX 3: <br> AGEING STRUCTURES OF JUVENILE LAKE STURGEON CAUGHT IN THE KEEYASK STUDY AREA.

Figure A3-1: Ageing structure from a wild juvenile Lake Sturgeon (8-year-old) caught in
Gull Lake. ..... 98

Figure A3-2: Ageing structure from a hatchery reared juvenile Lake Sturgeon caught in Stephens Lake ( 2 -year-old). Agers noted the presence of a weak first annulus and false annuli typically observed in hatchery reared Lake Surgeon99


Figure A3-1: Ageing structure from a wild juvenile Lake Sturgeon (8-year-old) caught in Gull Lake.


Figure A3-2: Ageing structure from a hatchery reared juvenile Lake Sturgeon caught in Stephens Lake (2-year-old). Agers noted the presence of a weak first annulus and false annuli typically observed in hatchery reared Lake Surgeon.

# APPENDIX 4: <br> WILD AND HATCHERY LAKE STURGEON RECAPTURE DATA, FALL 2018. 

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018.101
Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018 ..... 106

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018.

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Split Lake | 79284 | 900226000153736 | SPL-A | 12-Sep-18 | 683 | 779 | 2200 | 7 | 3.76 | 1121 |
| Split Lake | - | - | SPL-A | 18-Aug-15 | 507 | 579 | 820 | - | - | - |
|  |  |  | Growth |  | 176 | 200 | 1380 |  |  |  |
| Split Lake | 91364 | 900226000548175 | SPL-A | 12-Sep-18 | 996 | 1160 | 7350 | - | 3.36 | 1912 |
| Split Lake | - | - | SPL-A | 18-Jun-13 | 859 | 962 | 5670 | - | - | - |
|  |  |  | Growth |  | 137 | 198 | 1680 |  |  |  |
| Split Lake | 91600 | 900226000153775 | SPL-A | 12-Sep-18 | 796 | 892 | 4050 | 13 | 3.05 | 1854 |
| Split Lake | - | - | SPL-A | 15-Aug-13 | 580 | 652 | 1300 | - | - | - |
|  |  |  | Growth |  | 216 | 240 | 2750 |  |  |  |
| Burntwood River | 88179 | 900226000893378 | BWR-A | 13-Sep-18 | 749 | 822 | 2650 | - | 0.41 | 455 |
| Burntwood River | - | - | BWR-A | 15-Jun-17 | 840 | 980 | 2903 | - | - | - |
|  |  |  |  |  | -91 ${ }^{\text {b }}$ | -158 ${ }^{\text {b }}$ | -253 ${ }^{\text {b }}$ |  |  |  |
| Future Keeyask Reservoir | 103649 | 900226000629219 | GL-C | 11-Sep-18 | 873 | 984 | 5300 | - | 1.63 | 1529 |
| Future Keeyask Reservoir | - | - | GL-C | 05-Jul-14 | 764 | 844 | 3600 | - | - | - |
|  |  |  | Growth |  | 109 | 140 | 1700 |  |  |  |
| Future Keeyask Reservoir | 106462 | 900226000893868 | GL-C | 11-Sep-18 | 707 | 797 | 2450 | $10^{\text {a }}$ | 0.15 | 367 |
| Future Keeyask Reservoir | - | - | GL-C | 09-Sep-17 | 700 | 788 | 2500 | 9 | - | - |
|  |  |  | Growth |  | 7 | 9 | $-50{ }^{\text {b }}$ |  |  |  |
| Future Keeyask Reservoir | 107115 | 900226000768485 | GL-B | 12-Sep-18 | 955 | 1084 | 7100 | - | 7.47 | 819 |
| Future Keeyask Reservoir | - | - | GL-A | 15-Jun-16 | 900 | 1100 | 6350 | - | - | - |
|  |  |  | Growth |  | 55 | -16 ${ }^{\text {b }}$ | 750 |  |  |  |
| Future Keeyask Reservoir | 103116 | 900226000768467 | GL-C | 12-Sep-18 | 767 | 871 | 3200 | 9 | 0.38 | 828 |
| Future Keeyask Reservoir | - | - | GL-C | 06-Jun-16 | 620 | 711 | 2722 | - | - | - |
| Future Keeyask Reservoir | - | - | GL-C | 28-Jul-13 | 483 | 555 | 700 | - | - | - |
|  |  |  | Growth |  | 284 | 316 | 2500 |  |  |  |

Aquatic Effects Monitoring Plan

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018 (continued).

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | 89658 | 900226000327546 | GL-C | 12-Sep-18 | 770 | 858 | 3100 | 10 | 0.27 | 2911 |
| Future Keeyask Reservoir | - | - | GL-C | 23-Sep-10 | 347 | 385 | 250 | - | - | - |
|  |  |  | Growth |  | 423 | 473 | 2850 |  |  |  |
| Future Keeyask Reservoir | 94889 | 900226000327545 | GL-B | 13-Sep-18 | 664 | 750 | 1900 | 10 | 2.76 | 2546 |
| Future Keeyask Reservoir | - | - | GL-C | 24-Sep-11 | 392 | 445 | 350 | 3 | - | - |
|  |  |  | Growth |  | 272 | 305 | 1550 |  |  |  |
| Future Keeyask Reservoir | 113033 | 900067000121673 | GL-C | 13-Sep-18 | 332 | 372 | 250 | 3 | 2.44 | 360 |
| Future Keeyask Reservoir | - | - | GL-C | 18-Sep-17 | 289 | 316 | 150 | 2 | - | - |
|  |  |  | Growth |  | 43 | 56 | 100 |  |  |  |
| Future Keeyask Reservoir | 82647 | 900226000629140 | GL-C | 13-Sep-18 | 936 | 1049 | 7000 | - | 7.08 | 1532 |
| Future Keeyask Reservoir | - | - | GL-B | 04-Jul-14 | 827 | 911 | 4536 | - | - | - |
| Future Keeyask Reservoir | - | - | GL-B | 23-Aug-06 | 563 | 643 | 726 | - | - | - |
|  |  |  | Growth |  | 373 | 406 | 6274 |  |  |  |
| Future Keeyask Reservoir | 106469 | 900226000893798 | GL-C | 13-Sep-18 | 512 | 588 | 850 | 5 | 1.58 | 366 |
| Future Keeyask Reservoir | - | - | GL-C | 12-Sep-17 | 505 | 579 | 800 | 4 | - | - |
|  |  |  | Growth |  | 7 | 9 | 50 |  |  |  |
| Future Keeyask Reservoir | 113824 | 900226000768413 | GL-C | 15-Sep-18 | 708 | 801 | 2450 | 10 | 0.39 | 830 |
| Future Keeyask Reservoir | - | - | GL-C | 07-Jun-16 | 619 | 706 | 2268 | - | - | - |
|  |  |  | Growth |  | 89 | 95 | 182 |  |  |  |
| Future Keeyask Reservoir | 113821 | 900043000103674 | GL-A | 15-Sep-18 | 486 | 549 | 650 | 5 | 9.22 | 1525 |
| Future Keeyask Reservoir | - | - |  | 13-Jul-14 | 242 | 272 | 100 | - | - | - |
|  |  |  | Growth |  | 244 | 277 | 550 |  |  |  |
| Future Keeyask Reservoir | 106463 | 900226000154225 | GL-C | 15-Sep-18 | 723 | 824 | 2850 | 10 | 1.77 | 371 |
| Future Keeyask Reservoir | - | - | GL-C | 09-Sep-17 | 700 | 800 | 2850 | 9 | - | - |
|  |  |  | Growth |  | 23 | 24 | 0 |  |  |  |

Aquatic Effects Monitoring Plan

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018 (continued).

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | 113814 | 900226000629005 | GL-C | 15-Sep-18 | 675 | 764 | 2450 | $9^{\text {a }}$ | 2.19 | 1543 |
| Future Keeyask Reservoir | - | - |  | 25-Jun-14 | 496 | 570 | 1000 | 6 | - | - |
|  |  |  | Growth |  | 179 | 194 | 1450 |  |  |  |
| Future Keeyask Reservoir | 112543 | 900226000629452 | GL-A | 16-Sep-18 | 574 | 651 | 1100 | 7* | 5.77 | 1094 |
| Future Keeyask Reservoir | 103463 | - | GL-B | 18-Sep-15 | 446 | 512 | 539 | $4^{\text {d }}$ | - | - |
| Future Keeyask Reservoir | 103463 | - | GL-B | 11-Sep-14 | 399 | 457 | 425 | 3 | - | - |
|  |  |  | Growth |  | 175 | 194 | 675 |  |  |  |
| Future Keeyask Reservoir | 105691 | 900043000103824 | GL-B | 17-Sep-18 | 668 | 766 | 1850 | $10^{*}$ | 0.30 | 1095 |
| Future Keeyask Reservoir | - | - | GL-B | 18-Sep-15 | 574 | 661 | 1100 | 7 | - | - |
| Future Keeyask Reservoir | - | - | GL-B | 02-Jul-14 | 510 | 595 | 1150 | - | - | - |
|  |  |  | Growth |  | 158 | 171 | 700 |  |  |  |
| Future Keeyask Reservoir | 112502 | 900226000893677 | GL-C | 17-Sep-18 | 669 | 765 | 2300 | $10^{\text {a }}$ | 1.72 | 734 |
| Future Keeyask Reservoir | - | - | GL-C | 13-Sep-16 | 622 | 710 | 2120 | 7 | - | - |
|  |  |  | Growth |  | 47 | 55 | 180 |  |  |  |
| Future Keeyask Reservoir | 113841 | 900043000103668 | GL-B | 18-Sep-18 | 489 | 556 | 700 | 5 | 4.59 | 1465 |
| Future Keeyask Reservoir | - | - | GL-C | 14-Sep-14 | 237 | 271 | 50 | 1 | - | - |
|  |  |  | Growth |  | 252 | 285 | 650 |  |  |  |
| Future Keeyask Reservoir | 91383 | 900226000629177 | GL-C | 19-Sep-18 | 1031 | 1112 | 8500 | - | 21.21 | 1546 |
| Future Keeyask Reservoir | - | - | BR-D | 26-Jun-14 | 895 | 1003 | 5897 | - | - | - |
|  |  |  | Growth |  | 136 | 109 | 2603 |  |  |  |
| Stephens Lake | 101500 | 900226000628170 | STL-B | 12-Sep-18 | 812 | 909 | 3650 | $10^{*}$ | 0.68 | 1076 |
| Stephens Lake | - | - | STL-B | 02-Oct-15 | 700 | 763 | 2400 | 7 | - | - |
|  |  |  | Growth |  | 112 | 146 | 1250 |  |  |  |
| Stephens Lake | 110580 | 900067000121406 | STL-A | 12-Sep-18 | 386 | 435 | 300 | 3 | 2.83 | 725 |
| Stephens Lake | - | - | STL-B | 17-Sep-16 | 247 | 280 | 80 | 1 | - | - |
|  |  |  | Growth |  | 139 | 155 | 220 |  |  |  |

Aquatic Effects Monitoring Plan
Juvenile Lake Sturgeon Population

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018 (continued).

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | 115826 | 900226000154078 | STL-A | 13-Sep-18 | 837 | 945 | 4300 | - | 0.66 | 94 |
| Stephens Lake | - | - | STL-A | 11-Jun-18 | 825 | 925 | 4275 | - | - | - |
|  |  |  |  |  | 12 | 20 | 25 |  |  |  |
| Stephens Lake | 110994 | 900226000548945 | STL-B | 15-Sep-18 | 522 | 595 | 1100 | 5 | 1.10 | 825 |
| Stephens Lake | - | - | STL-A | 12-Jun-16 | 343 | 387 | 253 | - | - | - |
|  |  |  | Growth |  | 179 | 208 | 847 |  |  |  |
| Stephens Lake | 112945 | 900226000893689 | STL-B | 15-Sep-18 | 545 | 630 | 1150 | 5 | 0.30 | 359 |
| Stephens Lake | - | - | STL-B | 21-Sep-17 | 513 | 589 | 1000 | 4 | - | - |
|  |  |  | Growth |  | 32 | 41 | 150 |  |  |  |
| Stephens Lake | 92093 | N/A | STL-B | 15-Sep-18 | 770 | 874 | 3300 | $10^{*}$ | 2.89 | 2191 |
| Stephens Lake | - | - | STL-A | 15-Sep-12 | 461 | 540 | 700 | 4 | - | - |
| Stephens Lake | - | - | STL-B | 06-Jul-11 | 340 | 398 | 300 | - | - | - |
|  |  |  | Growth |  | 430 | 476 | 3000 |  |  |  |
| Stephens Lake | 110582 | 900067000055566 | STL-B | 15-Sep-18 | 530 | 616 | 1050 | $5^{\text {a }}$ | 3.06 | 728 |
| Stephens Lake | - | - | STL-B | 17-Sep-16 | 416 | 471 | 440 | 2 | - | - |
|  |  |  | Growth |  | 114 | 145 | 610 |  |  |  |
| Stephens Lake | 96513 | 900226000154249 | STL-B | 16-Sep-18 | 531 | 602 | 1075 | $6{ }^{*}$ | 15.01 | 1095 |
| Stephens Lake | - | - | STL-A | 12-Sep-17 | 521 | 585 | 950 | $5{ }^{*}$ | - | - |
| Future Keeyask Reservoir | - | $900043000103672^{\text {c }}$ | GL-C | 17-Sep-15 | 382 | 435 | 363 | 3 | - | - |
| Future Keeyask Reservoir | - | - | GL-C | 16-Sep-14 | 343 | 392 | 300 | 2 | - | - |
|  |  |  | Growth |  | 188 | 210 | 775 |  |  |  |
| Stephens Lake | 111064 | 900067000121333 | STL-B | 18-Sep-18 | 385 | 430 | 375 | $4^{\text {a }}$ | 2.47 | 370 |
| Stephens Lake | - | - | STL-A | 13-Sep-17 | 349 | 387 | 250 | 2 | - | - |
|  |  |  | Growth |  | 36 | 43 | 125 |  |  |  |

Table A4-1: Original capture date and biological data for fish recaptured in gill nets, fall 2018 (continued).

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | 101994 | 9001380344626 | STL-B | 19-Sep-18 | 430 | 477 | 600 | $4^{\text {a }}$ | 0.49 | 1090 |
| Stephens Lake | - | - | STL-B | 25-Sep-15 | 230 | 246 | 180 | 1 | - | - |
|  |  |  | Growth |  | 200 | 231 | 420 |  |  |  |
| Stephens Lake | 112924 | 900226000768894 | STL-B | 20-Sep-18 | 410 | 447 | 550 | 3 | 2.33 | 371 |
| Stephens Lake | - | - | STL-A | 14-Sep-17 | 363 | 398 | 280 | 2 | - | - |
|  |  |  |  |  | 47 | 49 | 270 |  |  |  |

*     - Ages assigned based on structures aged in a previous study year.
a - Assigned age in 2018 differed from ageing carried out in a previous year.
b - Measurement discrepancies due to errors in measurement at release or recapture.
c - Original PIT tag has been lost.
d - Aged as 3 years old in both 2014 and 2015.

Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018.

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Split Lake | 113710 | 900043000102957 | SPL-A | 20-Sep-18 | 520 | 581 | 1100 | 5 | 32.29 | 1449 |
| Burntwood River | - | - | BWR-B | 02-Oct-14 | 265 | 297 | 105 |  |  |  |
|  |  |  | Growth |  | 255 | 284 | 995 |  |  |  |
| Future Keeyask Reservoir | 113006 | 900067000112175 | GL-B | 11-Sep-18 | 353 | 409 | 300 | 2 | 8.79 | 460 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 216 | 263 | 59 |  |  |  |
|  |  |  | Growth |  | 137 | 146 | 241 |  |  |  |
| Future Keeyask Reservoir | 113022 | 900067000059369 | GL-A | 13-Sep-18 | 349 | 390 | 250 | 2 | 2.63 | 462 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 220 | 250 | 63 |  |  |  |
|  |  |  | Growth |  | 129 | 140 | 187 |  |  |  |
| Future Keeyask Reservoir | 113023 | 900067000059421 | GL-A | 13-Sep-18 | 365 | 419 | 300 | 2 | 2.63 | 462 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 232 | 271 | 74 |  |  |  |
|  |  |  | Growth |  | 133 | 148 | 226 |  |  |  |
| Future Keeyask Reservoir | 113029 | 900067000055044 | GL-B | 13-Sep-18 | 428 | 497 | 450 | 4 | 2.72 | 1179 |
| Future Keeyask Reservoir | - | - | GL-B | 22-Jun-15 | 209 | 247 | 53 |  |  |  |
|  |  |  | Growth |  | 219 | 250 | 397 |  |  |  |
| Future Keeyask Reservoir | 113044 | 900067000055461 | GL-C | 14-Sep-18 | 424 | 488 | 450 | 4 | 3.92 | 1094 |
| Future Keeyask Reservoir | - | - | GL-B | 16-Sep-15 | 295 | 343 | 136 |  |  |  |
|  |  |  | Growth |  | 129 | 145 | 314 |  |  |  |
| Future Keeyask Reservoir | 113048 | 900067000112161 | GL-B | 14-Sep-18 | 403 | 458 | 350 | 2 | 8.5 | 463 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 287 | 86 | - |  |  |  |
|  |  |  | Growth |  | 116 | 372 | - |  |  |  |
| Future Keeyask Reservoir | 113831 | 900067000055689 | GL-B | 17-Sep-18 | 440 | 511 | 500 | 4 | 0.48 | 1097 |
| Future Keeyask Reservoir | - | - | GL-B | 16-Sep-15 | 318 | 366 | 165 |  |  |  |
|  |  |  | Growth |  | 122 | 145 | 335 |  |  |  |

Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018 (continued).

| Location | Floy-tag \# | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | 113832 | 900067000055259 | GL-B | 17-Sep-18 | 405 | 466 | 400 | 4 | 0.48 | 1097 |
| Future Keeyask Reservoir | - | - | GL-B | 16-Sep-15 | 291 | 335 | 144 |  |  |  |
|  |  |  | Growth |  | 114 | 131 | 256 |  |  |  |
| Future Keeyask Reservoir | 113833 | 900067000112140 | GL-B | 17-Sep-18 | 373 | 428 | 300 | 2 | 9.16 | 466 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 330 | 128 | - |  |  |  |
|  |  |  | Growth |  | 43 | 300 | - |  |  |  |
| Future Keeyask Reservoir | 113835 | 900067000058508 | GL-B | 17-Sep-18 | 453 | 519 | 500 | 4 | 2.80 | 1183 |
| Future Keeyask Reservoir | - | - | GL-B | 22-Jun-15 | 215 | 249 | 60 |  |  |  |
|  |  |  | Growth |  | 238 | 270 | 440 |  |  |  |
| Future Keeyask Reservoir | 113836 | 900067000112127 | GL-B | 17-Sep-18 | 400 | 463 | 400 | 2 | 8.92 | 466 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 233 | 277 | 72 |  |  |  |
|  |  |  | Growth |  | 167 | 186 | 328 |  |  |  |
| Future Keeyask Reservoir | 113839 | 900067000059354 | GL-A | 18-Sep-18 | 318 | 365 | 200 | 2 | 2.27 | 467 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 228 | 266 | 71 |  |  |  |
|  |  |  | Growth |  | 90 | 99 | 129 |  |  |  |
| Future Keeyask Reservoir | 113848 | 900067000056730 | GL-C | 18-Sep-18 | 381 | 438 | 400 | 2 | 12.8 | 467 |
| Future Keeyask Reservoir | - | - | GL-A | 08-Jun-17 | 249 | 290 | 90 |  |  |  |
|  |  |  | Growth |  | 132 | 148 | 310 |  |  |  |
| Future Keeyask Reservoir | 113153 | 900067000058648 | GL-B | 19-Sep-18 | 479 | 545 | 700 | 4 | 0.39 | 1185 |
| Future Keeyask Reservoir | - | - | GL-B | 22-Jun-15 | 232 | 271 | 75 |  |  |  |
|  |  |  | Growth |  | 247 | 274 | 625 |  |  |  |
| Future Keeyask Reservoir | 113154 | 900067000055405 | GL-B | 19-Sep-18 | 462 | 535 | 600 | 4 | 0.45 | 1099 |
| Future Keeyask Reservoir |  |  | GL-B | 16-Sep-15 | 297 | 344 | 130 |  |  |  |
|  |  |  | Growth |  | 165 | 191 | 470 |  |  |  |

Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018 (continued).

| Location | $\begin{aligned} & \text { Floy-tag } \\ & \quad \# \end{aligned}$ | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Future Keeyask Reservoir | 113161 | 900067000055447 | GL-B | 19-Sep-18 | 448 | 518 | 500 | 4 | 2.78 | 1185 |
| Future Keeyask Reservoir | - | - | GL-B | 22-Jun-15 | 257 | 298 | 93 |  |  |  |
|  |  |  | Growth |  | 191 | 220 | 407 |  |  |  |
| Future Keeyask Reservoir | -9999 | 900067000110427 | GL-C | 19-Sep-18 | 255 | 296 | 100 | 1 | 127.8 | 104 |
| Burntwood River | - | - | BWR | 07-Jun-18 | 193 | 228 | 47 |  |  |  |
|  |  |  | Growth |  | 62 | 68 | 53 |  |  |  |
| Stephens Lake | 100673 | 900067000055481 | STL-A | 11-Sep-18 | 455 | 525 | 750 | 4 | 11.19 | 1091 |
| Future Keeyask Reservoir | - | - | GL-C | 16-Sep-15 | 295 | 349 | 140 |  |  |  |
|  |  |  | Growth |  | 160 | 176 | 610 |  |  |  |
| Stephens Lake | 115144 | 900067000058474 | STL-A | 11-Sep-18 | 490 | 562 | 875 | 4 | 0.66 | 1093 |
| Stephens Lake | - | - | STL-A | 14-Sep-15 | 277 | 322 | 118 |  |  |  |
|  |  |  | Growth |  | 213 | 240 | 757 |  |  |  |
| Stephens Lake | 113259 | 900067000055178 | STL-B | 15-Sep-18 | 503 | 570 | 900 | 4 | 1.83 | 1181 |
| Stephens Lake | - | - | STL-B | 22-Jun-15 | 219 | 250 | 54 |  |  |  |
|  |  |  | Growth |  | 284 | 320 | 846 |  |  |  |
| Stephens Lake | 113261 | 900067000055526 | STL-A | 15-Sep-18 | 495 | 566 | 850 | 4 | 2.34 | 1181 |
| Stephens Lake | - | - | STL-B | 22-Jun-15 | 221 | 258 | 57 |  |  |  |
|  |  |  | Growth |  | 274 | 308 | 793 |  |  |  |
| Stephens Lake | 113263 | 900067000055507 | STL-A | 15-Sep-18 | 497 | 566 | 875 | 4 | 3.29 | 1097 |
| Stephens Lake | - | - | STL-B | 14-Sep-15 | 314 | 360 | 180 |  |  |  |
|  |  |  | Growth |  | 183 | 206 | 695 |  |  |  |
| Stephens Lake | 113264 | 900067000055170 | STL-A | 15-Sep-18 | 480 | 550 | 775 | 4 | 13.14 | 1095 |
| Future Keeyask Reservoir | - | - | GL-B | 16-Sep-15 | 350 | 400 | 248 |  |  |  |
|  |  |  | Growth |  | 130 | 150 | 527 |  |  |  |

Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018 (continued).

| Location | $\begin{gathered} \text { Floy-tag } \\ \quad \# \end{gathered}$ | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | 113265 | 900067000111882 | STL-A | 15-Sep-18 | 354 | 410 | 375 | 2 | 1.7 | 457 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 218 | 255 | 62 |  |  |  |
|  |  |  | Growth |  | 136 | 155 | 313 |  |  |  |
| Stephens Lake | 113269 | 900067000055076 | STL-B | 17-Sep-18 | 502 | 568 | 700 | 4 | 0.91 | 1099 |
| Stephens Lake | - | - | STL-B | 14-Sep-15 | 298 | 344 | 140 |  |  |  |
|  |  |  | Growth |  | 204 | 224 | 560 |  |  |  |
| Stephens Lake | 113271 | 900067000111990 | STL-B | 17-Sep-18 | 375 | 416 | 350 | 2 | 4 | 459 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 262 | 300 | 103 |  |  |  |
|  |  |  | Growth |  | 113 | 116 | 247 |  |  |  |
| Stephens Lake | 113285 | 900067000113506 | STL-B | 19-Sep-18 | 392 | 405 | 425 | 2 | 3.23 | 461 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 240 | 281 | 79 |  |  |  |
|  |  |  | Growth |  | 152 | 124 | 346 |  |  |  |
| Stephens Lake | 113293 | 900067000058406 | STL-B | 20-Sep-18 | 470 | 535 | 700 | 4 | 14.99 | 1186 |
| Future Keeyask Reservoir | - | - | GL-B | 22-Jun-15 | 206 | 239 | 52 |  |  |  |
|  |  |  | Growth |  | 264 | 296 | 648 |  |  |  |
| Stephens Lake | 113295 | 900067000059240 | STL-B | 20-Sep-18 | 350 | 402 | 300 | 2 | 3.14 | 462 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 230 | 270 | 76 |  |  |  |
|  |  |  | Growth |  | 120 | 132 | 224 |  |  |  |
| Stephens Lake | 100674 | 900067000112046 | STL-B | 20-Sep-18 | 357 | 411 | 300 | 2 | 2.8 | 462 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 230 | 286 | 73 |  |  |  |
|  |  |  | Growth |  | 127 | 125 | 227 |  |  |  |
| Stephens Lake | 113296 | 900067000055532 | STL-B | 20-Sep-18 | 472 | 532 | 675 | 4 | 0.66 | 1102 |
| Stephens Lake | - | - | STL-A | 14-Sep-15 | 288 | 335 | 137 |  |  |  |
|  |  |  | Growth |  | 184 | 197 | 538 |  |  |  |

Table A4-2: Original release date and biological data for hatchery-reared Lake Sturgeon captured in gill nets, fall 2018 (continued).

| Location | $\begin{aligned} & \text { Floy-tag } \\ & \# \end{aligned}$ | Pit-tag No. | Zone | Date | Fork Length (mm) | Total Length (mm) | Weight (g) | Age | Distance (km) | Days Between Capture |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stephens Lake | 113297 | 900067000112569 | STL-B | 20-Sep-18 | 346 | 396 | 275 | 2 | 2.8 | 462 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 210 | 245 | 55 |  |  |  |
|  |  |  | Growth |  | 136 | 151 | 220 |  |  |  |
| Stephens Lake | 113300 | 900067000112320 | STL-B | 20-Sep-18 | 346 | 403 | 300 | 2 | 2.8 | 462 |
| Stephens Lake | - | - | STL-A | 15-Jun-17 | 232 | 273 | 76 |  |  |  |
|  |  |  | Growth |  | 114 | 130 | 224 |  |  |  |
| Stephens Lake | 113283 | 900067000055210 | STL-B | 20-Sep-18 | 464 | 532 | 700 | 4 | 0.7 | 1102 |
| Stephens Lake | - | - | STL-A | 14-Sep-15 | 274 | 317 | 123 |  |  |  |
|  |  |  | Growth |  | 190 | 215 | 577 |  |  |  |

## APPENDIX 5: POPULATION ESTIMATE INFORMATION.

$\begin{array}{ll}\text { Table A5-1: } & \begin{array}{l}\text { Results of POPAN analysis of juvenile Lake Sturgeon from the future } \\ \\ \text { Keeyask reservoir. .................................................................................. } 114\end{array}\end{array}$
Table A5-2: Results of POPAN analysis of juvenile Lake Sturgeon from Stephens Lake. ... 115
Table A5-3: Results of POPAN analysis of hatchery-reared juvenile Lake Sturgeon from the future Keeyask reservoir (FKR) and Stephens Lake116

Mark-recapture population estimates were been calculated for wild fish in the future Keeyask reservoir and Stephens Lake during the fall of eight different years (2010 and 2012-2018). Only wild Lake Sturgeon classified as juveniles (i.e., fork length less than 800 mm ) were included in the population estimate.

Data were analysed using the program MARK (White and Burnham 1999; Kendall 2001; Arnason and Schwartz 2002), which is an industry standard for the analysis of data from marked populations. Program MARK uses binary numbers to represent the encounter history of individuals, and then uses the cumulative pattern of 0's (not-encountered) and 1's (reencountered live capture) to generate a probability distribution of tag recaptures which form the basis of population estimation. Re-encounters can also be from dead recoveries (e.g., the animal is harvested) in which case the model uses a value of -1 . For example, the history "10-1" indicates than an animal was captured for the first time at sampling occasion 1, not encountered at sampling occasion 2, and recovered dead at sampling occasion 3.

Several different population model variants exist, most of which can be classified as either closed or open models. Closed models assume there are no births, deaths, immigration, or emigration between sample periods, while open models assume these processes occur. The Jolly-Seber model (POPAN formulation; Arnason and Schwarz 2002), as implemented within MARK, was used to estimate the annual abundance of juvenile Lake Sturgeon. This is an open model that requires few assumptions and modeled variables, and thus likely provides a reliable estimate of abundance.

Using first-time capture and recapture information, POPAN estimates the survival (i.e., the probability that a fish will survive from one capture to the next), the probability of recapture ( $p$; i.e., the probability that a fish will be recaptured given that the animal is alive and in the study area), and abundance ( N ; i.e., the number of juvenile Lake Sturgeon in the area during each capture period) (Tables A5-1 and A5-2).

- Model fit for survival was calculated as $77 \%$ for the future Keeyask reservoir and $88 \%$ for Stephens Lake.
- The probability of recapture varied among years.
- Recapture rates were split into two groups based on the model for the future Keeyask reservoir: i) 2010 and 2013 (0.02) and ii) 2012 and 2014-2018 (0.03).
- For Stephens Lake, recapture rates were split into four groups: i) 2010 and 2013 (0.03); ii) 2015 and 2018 (0.05); iii) 2014 (0.06); and iv) 2012, 2016, and 2017 (0.09).
- Abundance estimates for both the future Keeyask reservoir and Stephens Lake are provided for the 2010 and 2012-2018 study years.
- As sampling continues (i.e., year to year) and data is added to the model, the parameters are recalculated. Thus, although survival rates and abundance estimates are calculated for the same time periods, they may differ among reporting periods. This allows the estimates to become more refined and precise over time.

The Cormack-Jolly-Seber model was used to calculate an estimate of survival of hatcheryreared lake Sturgeon in both the future Keeyask reservoir and Stephens Lake between 2015 (when stocking began) and 2018. This model calculates an estimate using the probability of recapture. The probability of recapture differed each year, but were the same for both the future Keeyask reservoir and Stephens Lake (Table A5-2). A survival estimate could only be calculated for fish stocked in 2015 (i.e., 2014 cohort) due to low levels of recapture for fish stocked in other years.

## References

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Kendall, W.L. 2001. The robust design for capture-recapture studies: Analysis using Program MARK. In Wildlife, Land, and People: Priorities for the $21^{\text {st }}$ Century. Proceedings of the Second International Wildlife Management Congress. Edited by R. Field, R.J. Warren, H. Okarma, and P.R. Sievert. The Wildlife Society, Bethesda, Maryland, USA. p. 350-356.

White, G.C. and Burnham, K.P. 1999. Program MARK: Survival estimation from populations of marked animals. Bird Study 46 Supplement: 120-138.

Table A5-1: Results of POPAN analysis of juvenile Lake Sturgeon from the future Keeyask reservoir. Best model was constant survival and variable recapture. Confidence intervals are rounded.

| Parameter | Mean | SE | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | 0.77 | 0.05 | Low |

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Table A5-2: Results of POPAN analysis of juvenile Lake Sturgeon from Stephens Lake. Best model was constant survival and variable recapture. Confidence intervals are rounded.

| Parameter | Mean | SE | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low | High |
| Survival (all years) | 0.88 | 0.04 | 0.78 | 0.94 |
| 2010, 2013 Recapture | 0.03 | 0.00 | 0.02 | 0.04 |
| 2015, 2018 Recapture | 0.05 | 0.01 | 0.04 | 0.07 |
| 2014 Recapture | 0.06 | 0.02 | 0.04 | 0.11 |
| 2012, 2016, 2017 Recapture | 0.09 | 0.01 | 0.06 | 0.12 |
| 2010 Abundance | 1152 | 159 | 880 | 1509 |
| 2012 Abundance | 924 | 146 | 679 | 1257 |
| 2013 Abundance | 812 | 147 | 570 | 1155 |
| 2014 Abundance | 713 | 151 | 473 | 1076 |
| 2015 Abundance | 835 | 128 | 619 | 1126 |
| 2016 Abundance | 734 | 131 | 519 | 1037 |
| 2017 Abundance | 1035 | 185 | 730 | 1466 |
| 2018 Abundance | 1101 | 219 | 749 | 1620 |

Table A5-3: Results of POPAN analysis of hatchery-reared juvenile Lake Sturgeon from the future Keeyask reservoir (FKR) and Stephens Lake. Best model was constant survival and variable recapture. Confidence intervals are rounded.

| Parameter | Mean | SE | $95 \%$ Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low | High |
| FKR Survival | 0.95 | 0.26 | 0.00 | 1.00 |
| Stephens Survival | 0.86 | 0.24 | 0.10 | 1.00 |
| 2015 Recapture | 0.01 | 0.00 | 0.00 | 0.02 |
| 2016 Recapture | 0.02 | 0.01 | 0.01 | 0.06 |
| 2017 Recapture | 0.05 | 0.04 | 0.01 | 0.22 |
| 2018 Recapture | 0.03 | 0.04 | 0.00 | 0.25 |
| FKR 2014 Cohort (stocked 2015) | 423 | - | - | - |
| FKR 2015 | 402 | - | - | - |
| FKR 2016 | - | - | - |  |
| FKR 2017 | - | - | - |  |
| FKR 2018 | 382 | - | - | - |
| Stephens 2014 Cohort (stocked 2015) | 418 | - | - | - |
| Stephens 2015 | 361 | - | - | - |
| Stephens 2016 | - | - | - |  |
| Stephens 2017 | - | - | - |  |

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[^0]:    ${ }^{1}$ Because this analysis includes a compilation of data from all sampling years, multiple captures of the same fish in different study years were included. Fish recaptured in the same sampling year were excluded.

[^1]:    a - Ages assigned based on structures aged in a previous study year.

